ACCEPTED MANUSCRIPT • OPEN ACCESS

Quantifying the potential for climate change mitigation of consumption options

To cite this article before publication: Diana Ivanova et al 2020 Environ. Res. Lett. in press https://doi.org/10.1088/1748-9326/ab8589

Manuscript version: Accepted Manuscript

Accepted Manuscript is "the version of the article accepted for publication including all changes made as a result of the peer review process, and which may also include the addition to the article by IOP Publishing of a header, an article ID, a cover sheet and/or an 'Accepted Manuscript' watermark, but excluding any other editing, typesetting or other changes made by IOP Publishing and/or its licensors"

This Accepted Manuscript is © 2020 The Author(s). Published by IOP Publishing Ltd.

As the Version of Record of this article is going to be / has been published on a gold open access basis under a CC BY 3.0 licence, this Accepted Manuscript is available for reuse under a CC BY 3.0 licence immediately.

Everyone is permitted to use all or part of the original content in this article, provided that they adhere to all the terms of the licence <u>https://creativecommons.org/licences/by/3.0</u>

Although reasonable endeavours have been taken to obtain all necessary permissions from third parties to include their copyrighted content within this article, their full citation and copyright line may not be present in this Accepted Manuscript version. Before using any content from this article, please refer to the Version of Record on IOPscience once published for full citation and copyright details, as permissions may be required. All third party content is fully copyright protected and is not published on a gold open access basis under a CC BY licence, unless that is specifically stated in the figure caption in the Version of Record.

View the article online for updates and enhancements.

Review of reviews

"Focus on demand-side solutions for transitioning to low-carbon societies"

Quantifying the potential for climate change mitigation of consumption options
 Diana Ivanova¹*, John Barrett¹, Dominik Wiedenhofer², Biljana Macura³, Max Callaghan⁴, Felix
 Creutzig⁴

- 11 6 ¹ School of Earth and Environment, University of Leeds, Leeds, United Kingdom
 - 7 ² Institute of Social Ecology, University of Natural Resources and Life Sciences, Vienna, Austria
 - 8 ³ Stockholm Environment Institute, Linnégatan 87D, Box 24218, 10451 Stockholm, Sweden
 - ⁹ ⁴ Mercator Research Institute on Global Commons and Climate Change, Berlin, Germany
- 17
 10
 *Corresponding author, <u>d.ivanova@leeds.ac.uk</u>

20 11 Abstract

21 12 Background22

Around two-thirds of global GHG emissions are directly and indirectly linked to household consumption, with a global average of about 6 tCO₂eq/cap. The average per capita carbon footprint of North America and Europe amount to 13.4 and 7.5 tCO₂eq/cap, respectively, while that of Africa and the Middle East - to 1.7 tCO₂eq/cap on average. Changes in consumption patterns to low-carbon alternatives therefore present a great and urgently required potential for emission reductions. In this paper, we synthesize emission mitigation potentials across the consumption domains of food, housing, transport and other consumption.

32 20 Methods

We systematically screened 6,990 records in the Web of Science Core Collections and Scopus. Searches were restricted to 1) reviews of lifecycle assessment studies and 2) multiregional input-output studies of household consumption, published after 2011 in English. We selected against pre-determined eligibility criteria and quantitatively synthesized findings from 53 studies in a meta-review. We identified 771 original options, which we summarized and presented in 61 consumption options with a positive mitigation potential. We used a fixed-effects model to explore the role of contextual factors (geographical, technical and socio-demographic factors) for the outcome variable (mitigation potential per capita) within consumption options.

Results and discussion

We establish consumption options with a high mitigation potential measured in tons of CO₂eq/capita/yr. For transport, the options with the highest mitigation potential include living car-free, shifting to a battery electric vehicle, and reducing flying by a long return flight with a median reduction potential of more than 1.7 tCO₂eq/cap. In the context of food, the highest carbon savings come from dietary changes, particularly an adoption of vegan diet with an average and median mitigation potential of 0.9 and 0.8 tCO₂eq/cap, respectively. Shifting to renewable electricity and refurbishment and renovation are the options with the highest mitigation potential in the housing domain, with medians at 1.6 and 0.9 tCO₂eq/cap, respectively. We find that the top 10 consumption options together yield an average mitigation potential of 9.2 tCO₂eq/cap, indicating substantial contributions towards achieving the 1.5-2°C target, particularly in high-income context.

Background

The need for demand reductions

Global greenhouse gas (GHG) emissions (carbon footprints) have been steadily rising, with faster, sizable and immediate CO₂ emissions declines needed to limit cumulative emissions and reach net zero emissions in 2050^1 . Annual GHG emissions must decrease by 45% percent of their 2010-levels by 2030, and reach net-zero by 2050 to limit temperature changes to 1.5°C above preindustrial levels. The potential impacts and risks are substantially lower for a 1.5° C global warming compared with a 2° C. including climate-related risks and threats regarding various ecosystems and human welfare¹. Global GHG emissions amounted to 6.3 tCO₂eq/cap in 2011²; however, these are highly unequally distributed across income groups and countries^{3–8}. For example, the average per capita carbon footprint of North America and Europe amount to 13.4 and 7.5 tCO₂eq/cap, respectively, while that of Africa and the Middle East - to 1.7 tCO₂eq/cap on average (SM figure 1). For a population of 8.5 billion by 2030⁹, emissions need to decrease to an average of $\sim 2.8 \text{ tCO}_2 \text{eq/cap}$ by 2030, to comply with a pathway of limiting climate change to 1.5°C of global warming. This is broadly in line with other estimates of per capita carbon budgets^{10–12}.

The exact carbon budget for limiting global warming to 1.5°C is influenced by uncertainty about earth system dynamics, as well as the scale and speed of adoption of negative emission technologies. Almost all of the IPCC scenarios currently assume large-scale adoption of negative emission technologies at massive scales^{13–15}, which are potentially associated with strong adverse economic and environmental consequences¹⁶, energy constraints (e.g. expanding carbon)¹⁷ and moral hazards because they tempt policy makers to delay mitigation action now¹⁵.

Energy end-use is the least efficient part of the global energy system with the largest improvement potential, where appropriate scaling down of the global energy demand allows for feasible de-carbonization without betting on controversial negative emission technologies or geoengineering. While technological solutions that decarbonize energy supply or capture carbon have to make a significant mitigation contribution, changing consumption offers more flexibility for reducing carbon intensity in the energy supply sector and limit the related supply-side risks¹⁸. Mitigation scenarios relying more heavily on reduction in the demand of energy services are clearly associated with the lowest mitigation and adaptation challenges^{15,19} and provide a range of co-benefits.

Challenging consumption

Behavior, everyday life and cultural norms around consumption have a crucial influence on energy use and embodied emissions, with a high mitigation potential in various consumption domains^{18,20,21}. 65% of global GHG emissions, and 50-80% of land, water and material use, can be directly and indirectly linked to household consumption³. Income is a major driver of household carbon footprints^{5,7,8,22,23}, directly affecting purchasing power of households. Changes in household consumption patterns to low-carbon alternatives, such as transport model shifts, home energy reduction and dietary shifts, thus present a great mitigation potential.

Importantly, in the last decade, so-called multiregional input-output models (MRIO) have enabled the systematic analysis of global production and consumption using consistent accounts of global GHG emissions, and taking into account the scale and complexity of international trade and supply chains²⁴⁻ ²⁶. Consumption estimates derived through MRIOs were the first to fully allocate global emissions to national household consumption (as well as government activities and investments) without double-counting or omitting emissions, thus overcoming a long-standing limitation of single-regional input-output approaches and lifecycle assessment (LCA) studies^{27,28}. However, understanding options for change also requires bottom-up detailed information and insights going down to the product-level – which is a challenge for MRIOs as they offer a quite limited product detail. In this context, LCAs are relevant due to their process-specific and highly detailed nature. Here we argue that a combination of

- bottom-up and top-down approaches provides a robust base for the review of the mitigation potentials of consumption options.
- In this paper, we systematically review the literature on mitigation potentials across various consumption
- domains, including food, housing, and transport, focusing on academic publications since 2011 to ensure
- relevance of derived estimates. While prior studies address some of these concerns (for a non-
- comprehensive list of studies see $^{11,16,29-31}$), we conduct meta-review including the more recent evidence.
- Therefore, we provide a richer and more updated evidence base to inform about mitigation potentials of
- changes in consumption practices, policies and infrastructure.
- For the purpose of this paper, we do not capture mitigation potential associated with other avenues towards social change²¹, such as community action and engagement^{32,33}, policies and incentives, political engagement and non-violent civil disobedience³⁴ or reductions in overall working time and re-definitions of paid labour²³, which all are highly relevant for challenging societal norms around consumption and tackling climate change. Supply chain actors play a key role for climate change mitigation, having direct agency over the majority of energy and emissions along supply chains^{35,36}. Similarly, structural change by governments, ending fossil-fuel support, and providing low-carbon infrastructures, is crucial to enable climate change mitigation^{37–39}. We also do not review system-wide effects and potential for income rebound effects⁴⁰⁻⁴². Our focus on consumption options should not be interpreted as passing the mitigation responsibility to consumers⁴³. Still, a change in consumption practices is needed for reaching net-zero carbon emissions^{1,44}.

Research questions

- Primary question: What is the mitigation potential of household-level consumption options within mobility, housing and food sectors, when considering GHG emissions along the whole lifecycle?
- The primary question consists of the following question components:
- Household consumption of food, mobility and housing *Population* (*P*):
- *Intervention (I)*: Consumption options within each end-use sector
- Average per capita carbon footprints of food, mobility and housing *Comparator* (*C*):
 - Annual carbon savings measured in per capita CO₂-equivalent reductions *Outcome* (*O*):
- LCA review studies with quantitative synthesis of data, MRIO studies of *Study types*: household consumption, consumption scenario studies
- We focus on household consumption associated with the three end-use sectors of food, transport and housing as they are highly relevant in terms of consumption-based GHG emissions^{3,45}, energy⁴⁶ and other resource use³ with some of the highest potential for consumption intervention^{29,47}.
- Secondary question: What factors may explain differences in carbon savings associated with each consumption option across studies and contexts?
- We aim to capture sources of heterogeneity across studies, including system boundary⁴⁸, methodological specificities, socio-economic, urban-rural and geographical context among others.

Methods and search results

- The review followed the Collaboration for Environmental Evidence Guidelines⁴⁹ and it conformed to ROSES reporting standards⁵⁰. It was conducted according to peer-reviewed protocol⁵¹ that was submitted to Environmental Research Letters in March 2019 and approved in April 2019. The approved protocol is openly available online⁵¹.
 - Deviations from the protocol (outline)

1 The following changes were made from the final published $protocol^{51}$: first, we applied machine learning

in the article screening process; second, we discussed the variation among studies in a qualitative manner
in text rather than using the CEESAT tool for critical assessment (which was not suitable to assess non-

4 review studies).

5 Searches for literature

6 Searches were performed on Web of Science Core Collections (WoSCC) and Scopus to identify relevant
7 peer-reviewed studies published after 2011, using the University of Leeds subscription. The searches
8 were done on titles, keywords and abstracts in English.

The search string was composed of three sub-strings: the GHG emission (X), study type (review) (Y) and consumption domain (Z) sub-string (Table 1). The sub-strings were connected with the Boolean operator "AND" as follows: X AND Y AND Z. We based the GHG emission sub-string (X) on prior similar searches^{52,53}. The consumption domain sub-string (Z) captured the consumption domains of transport, food, housing and other consumption (general), and specific consumption options (interventions) within these domains. The sub-strings in each domain-specific cell were connected with Boolean operator "OR" to form the consumption domain sub-string (Z). To test comprehensiveness of the search, we used a list of benchmark papers (see the protocol for details).

-								
	E	(((atmospheric OR ar	nthropogenic OR effect* OR emi	ssion* OR footprint* Ol	R mitigat* OR sav* OR			
X	sio	reduc* OR budget* OR impact* OR decreas*) AND (carbon OR CO2 OR CH4 OR methane OR N2O OR						
Sub-string X	GHG emission	nitrous oxide OR "gr	eenhouse gas*" OR GHG OR G	HGs)) OR (climat* AN	D (action* OR chang* OR			
str	j er	warm* OR shift*)) OR "global warming" OR "emission reduction*" OR (mitigation AND (action* OR						
-di	НC	potential*)) NOT (car	talyst* OR distill* OR chemicals	OR super-critical OR f	oaming OR pore OR			
$\mathbf{\tilde{s}}$	5	nanotube*))	(
		((lifecycle OR life-cy	cle OR "life cycle" OR LCA O	R embodied OR indired	ct OR embedded OR "supply			
Sub-string	a)		assessment*") AND (review* O					
iti.	ype	-	neta-stud* OR metastud* OR ov					
- d	dy 1	-	vs* OR aggrega*))) OR (((multire	-				
S S	Study type		put output")) OR MRIO))		6 /			
		(1) General	(2) Transport	(3) Food	(4) Housing			
		(consum* OR	((airplane* OR automobile*	(beef OR beverage*	("air condition*" OR			
		lifestyle* OR	OR bicycl* OR bik* OR	OR "calor* intake"	apartment* OR appliance*			
		demand* OR	bus* OR car* OR commut*	OR cereal* OR	OR boiler* OR cement OR			
		waste*)	OR cycl* OR *diesel OR	cheese OR chicken	clay OR concrete OR			
			driv* OR engine* OR flight*	OR dairy OR diet*	construct* OR cool* OR			
			OR fly* OR fuel* OR	OR egg* OR	dwelling* OR electronic*			
			gasoline OR "liquefied	fertilizer* OR fish	OR energy OR "floor			
			petroleum gas" OR LPG OR	OR food OR fruit*	space" OR heat* OR hemp			
	su		kerosene OR metro OR	OR grain* OR meat	OR home* OR hous* OR			
ш	nai		mobil* OR plane* OR ride*	OR milk OR plant*	light* OR "living space"			
Sub-string Z-term	dor		OR subway OR touris* OR	OR pork OR	OR metal* OR refrig* OR			
N	ouc		train* OR transit OR	restaurant OR sugar	rent* OR room OR sand			
ing	pti		transport* OR travel* OR	OR vegetable* OR	OR shelter OR "solar			
stri	Consumption domains		underground OR vehicle*))	yoghurt)	panel*" OR stone OR			
ģ	suc				timber OR window* OR			
S	Ŭ				"white good*" OR wood)			

AUTHOR SUBMITTED MANUSCRIPT - ERL-108137.R1

Page 5 of 24

1

1
2
3
4
5
6
7
/
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
34 35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

	(decreas* OR	("light weight" OR electric*	("eat less" OR	(cohous* OR co-hous* OR
	durab* OR eco*	OR hybrid* OR telecommut*	compost* OR	downsize* OR insulat* OR
	OR efficien* OR	OR telework* OR walk*)	flexitarian OR local	refurbish* OR renovat* OR
Consumption interventions	green* OR		OR organic OR	retrofit* OR ((temperature
	longetivity OR		season* OR vegan	OR thermal) AND
	natural OR		OR vegetarian)	(preference OR comfort OR
	maintain* OR			set-point* OR "set point*"
	recycl* OR reduc*			OR setting)))
	OR renewabl* OR			
	repair* OR reus*			
	OR "second hand"			
	OR second-hand			
	OR shar* OR			
Ŭ	sufficien*)			

Table 1: A summary of the sub-string X, Y and Z terms. The sub-strings are shown as formatted for Web of Science search..
 See the supplementary material for Scopus formatting.

A search on WoSCC (conducted on 24th May 2019) yielded 5,638 records and on Scopus additional
1,352 records (see the supplementary materials for search queries), totaling 6,990 records. The results
of both searches were combined into a "Scoping Review Helper" library where exact duplicates were
removed. Figure 1 provides more detailed overview of the search and screening process of the review.

7 Article screening and eligibility criteria

8 Article screening was done first at the title and abstract level, and then on full text level (Figure 1). The 9 title and abstract screening was supported by machine learning. Table 2 provides an overview the 10 eligibility criteria according to the PICO framework (see the supplementary information for more 11 details).

12 Having reviewed the first 991 records (15% of unique records) drawn randomly from the total number of records, we started an iterative process where at each iteration, we 1) trained a machine learning 13 14 model with the already screened documents; 2) fitted this model on the unseen documents; and 3) assigned the next set of documents for review by selecting the documents predicted to be most relevant. 15 16 We went through four iterations of machine learning prioritized screening, (see Figure 2.a) and each had 17 decreasing proportions of relevant documents in the set of reviewed records. The first iteration of 250 18 documents contained 38% of relevant records, while the last iteration of 100 documents - only 3% relevant documents. We screened a final random sample of 100 documents, and used this sample to 19 20 generate an estimate of the number of relevant documents remaining using the Agresti-Coull confidence interval. Figure 2.b) shows the minimum recall at different levels of uncertainty. 21

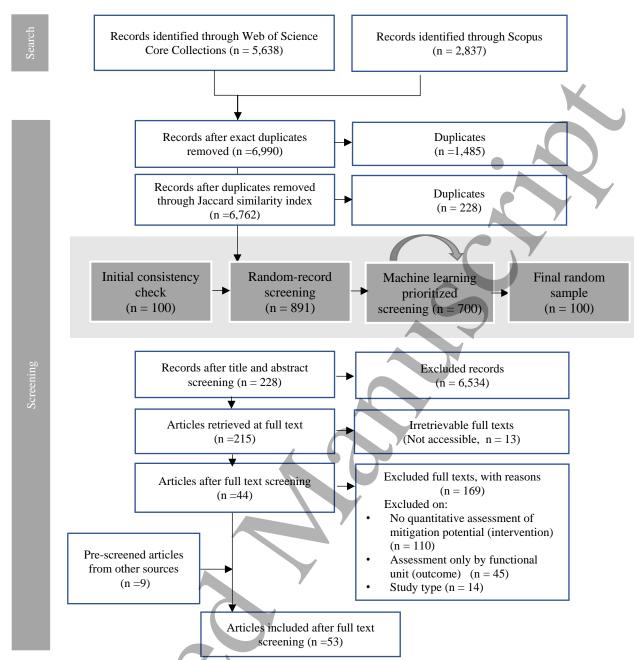


Figure 1: Flow diagram – adapted from the ROSES flow diagram for systematic reviews¹²⁸. See the supplementary data extraction for more detail about excluded articles and the supplementary materials for details on the methods.

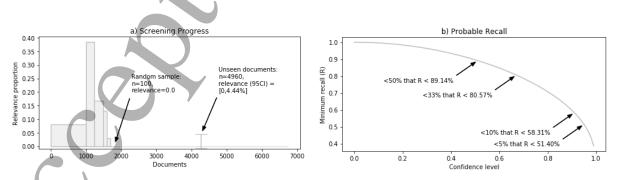


Figure 2: Screening progress (a) and probable recall (b). In a), each bar represents a set of screening decisions, with the width showing the number of documents and the height showing the percentage of them that was relevant. The first bar represents the 991 documents screened at random. The subsequent bars represent 4 sets of machine learning prioritised documents, a random sample of 100 documents, and remaining unseen documents. The random sample is used to generate the errorbar, the Agresti-Coull confidence interval. b) shows the probability distribution of the minimum level of recall, based on the assumption

	Inclusion criteria	Exclusion criteria
Eligible population/ setting	No geographical restriction and focus on household consumption	Mitigation potential not directly linked to households (e.g. government spending)
Eligible intervention: Consumption options by consumption domain	 Direct reduction -consumption reduction, shift between consumption categories, and curtailment. Examples include living car-free or avoiding flights (transport)²⁹, consuming fewer calories (food)⁵⁴ and conserve energy at home (housing)⁵⁵ Indirect reduction -changes in consumption patterns, changes in use behavior and changes in disposal patterns. Examples include carpooling (transport), sharing of food surplus (food), or equipment maintenance (housing)⁵⁶ Direct improvement -purchases of products that are more efficient in use or produced more efficiently. Examples include opting for electric vehicles (transport)⁵⁷, plant-based diet (food)^{29,54} and renewable energy (housing)²⁹. Indirect improvement -changes in disposal behavior. Examples include recycling batteries (transport), food packaging (food), electrical appliances (housing). 	Mitigation options beyond the adopted framework ⁵⁸ were out of scope. This includes macro-economic or industrial energy efficiency measures and technological solutions, producer incentives or other options on the supply side; population ¹¹ measures; mitigation potential of policies
Outcome: Mitigation potential and lifecycle emissions	Mitigation potential assessed through annual carbon savings in kilograms/tons CO ₂ -equivalents per capita, converting GHGs (e.g. CO ₂ , CH ₄ , N ₂ O, SF ₆) to equivalent amounts of CO ₂ (e.g. GWP100).	Focus only on direct emissions ⁵⁶ (e.g. well- to-wheel LCAs) or carbon intensities in functional units with no estimate of consumption; system-wide effects and potential for income rebound effects ^{40–42} . Consumption activities with high carbon intensity ^{3,59} should be considered to avoid rebound.
Study types	Supply chain lifecycle GHG emissions through LCA review studies and MRIO studies, physical trade flow or hybrid modelling studies, studies on re-designing of consumption.	Systematic maps and reviews with only narrative synthesis; mitigation assessment through regression coefficients.

1 Table 2: Eligibility criteria. See SI table 1 for more details on the inclusion and exclusion criteria.

After titles and abstract screening, we considered 228 relevant records at full-text (Figure 1). In addition, nine pre-screened articles were added separately, which were considered relevant but were not found through the original search. Six of these additions were not published at the time of the original search. We applied the inclusion and exclusion criteria (Table 2) and a final set of 53 articles were considered eligible at full text. See the supplementary materials and extraction sheet for more details on the procedure.

8 We used software for evidence synthesis "Scoping Review Helper" (developed by MCC Berlin), for 9 managing search results, removing duplicates, screening records, extracting data and conducting 10 synthesis. We also designed search queries through managing topics iteratively, and refined the 11 inclusion criteria during the screening process.

12 Data extraction and synthesis

We extracted meta-data from each reviewed study, including title, author team, year of publication and data collection, consumption option and domain, geographical context, method, system boundary, carbon metric and GHGs included from the eligible studies. We further extracted the study quantitative findings, e.g. average, standard deviation, number of studies reviewed, min-max range, absolute and relative carbon savings, contextual carbon footprint calculations. Missing or unclear information was requested directly from authors. We recalculated the mitigation potential of consumption options in tons CO_2 equivalents per capita where needed in order to improve comparability across studies.

The baselines considered in the reviewed studies are associated with large uncertainties and different
 assumptions (e.g. average baseline vs high-carbon baseline). At the same time, the baselines are key for

the calculation of mitigation potentials and may largely affect the order of consumption options on the
 graph. In such cases results should be interpreted with caution.

3 Data synthesis and potential effect modifiers/reasons for heterogeneity

Included literature is characterized by a large variation in methods, internal validity of studies, coverage of different GHGs, location and timeframe, system boundary, assumptions about uptake rate⁵⁶ and other potential sources of heterogeneity. We discussed heterogeneity along with the narrative synthesis of study findings. Where data allowed, we considered the effect modifiers in quantitative synthesis. We used a fixed-effects model to explore the relationship between predictors (various geographical, technical and socio-demographic factors) and outcome variables (mitigation potential per capita) across consumption options as a way to explain the variation in mitigation potential. Using the fixed-effect approach, we control for factors invariant across mitigation options, which we could not include directly in our model.

¹⁹₂₀ 13 Review Results

Figure 3-6 depict the mitigation potential ranges of various consumption options in the domains of food,
transport, housing and other consumption. Positive values are associated with positive mitigation
potential, with the options ordered by medians.

²⁵ 18 Transport

The highest mitigation potential of reviewed options is found in the domain of transport (Figure 3),
which is also associated with a substantial carbon footprint in most world regions (SM figure 1). The
consumption options with the highest mitigation potential advocate reduction in car and air travel, as
well as a shift toward less carbon intensive fuel sources, means and modes of transportation.

There is substantial mitigation potential in reducing air travel for those who fly. One less flight (long return) may reduce between 4.5 and 0.7 (mean of 1.9) tCO₂eq/cap, while taking One less flight (medium return) – between 1.5 and 0.2 (0.6) tCO₂eq/cap. The two options have a median reduction potential of 1.7 and 0.6 tCO₂eq/cap, respectively. Yet, the number of trips per passenger in 2018 amounted to 2.0 in the United States and to 3.6 and 4.8 in wealthy European countries such as Luxembourg and Norway, with the numbers projected to increase rapidly 60 . Other studies exploring partial reductions in air travel (Less transport by air) find an average reduction potential of 0.8 tCO₂eq/cap. The overall mitigation potentials strongly depend on income, as high-income households fly much more^{4,5,61}.

Reducing car travel is associated with substantial mitigation potential. Living car-free has the highest median mitigation potential across all of the reviewed options at 2.0 tCO₂eq/cap, with a range between 3.6 and 0.6 tCO₂eq/cap. Assumptions around vehicle and fuel characteristics as well as travel distance are key for the estimated mitigation potential, with the maximum value in our sample being associated with giving up an SUV²⁹. Partial car reductions, captured by the options of Less car transport, Shift to active transport and Shift to public transport in our sample, have an average mitigation potential between 0.6 and 1.0 tCO₂eq/cap. These options are generally limited to replacing short and urban car trips with alternative transportation modes or reducing leisure trips^{42,62–64}, which constitute a relatively small portion of all travel and its embodied emissions^{57,65,66}. Yet, active and public transport alternatives have much lower carbon intensities per travel km^{57,67,68}. Active and public transport are characterized by average carbon intensities at 0.00 and 0.09 kgCO₂eq/km, while individualized motorized transport at 0.23 kgCO₂eq/km⁵⁷. Telecommuting practices reduce commute emissions between 1.4 and 0.1 (mean of 0.4) tCO2eq/cap, while Car-pooling and car-sharing and Fuel efficient driving have an average carbon savings of 0.3 tCO₂eq/cap. The practice of ride-hailing, or receiving transportation from an unlicensed taxi service, may result in an increase in emissions as a result of "deadheading", the travelled miles without a passenger between hired rides⁶⁹. For example, a non-pooled ride-hailing trip generates 47% greater emissions per mile compared to a private car trip of an average fuel efficiency⁶⁹. The number of passenger sharing the trip makes a substantial difference in terms of mitigation potential, as well as the

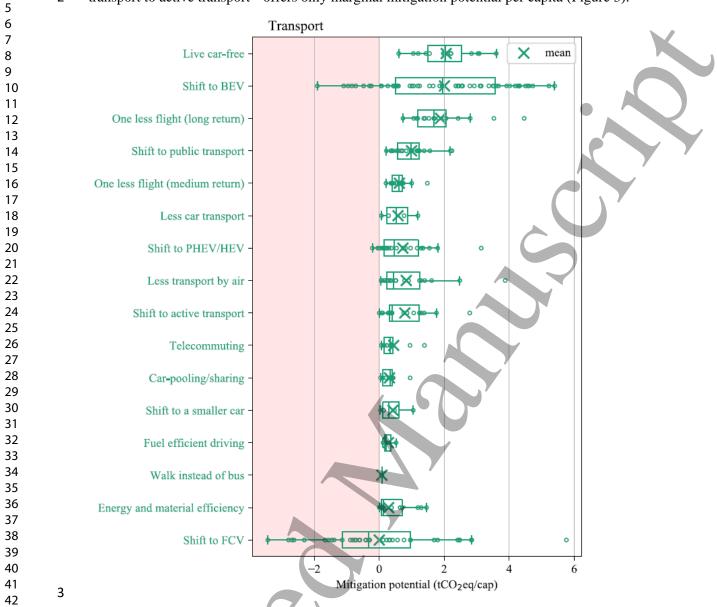


Figure 3: Annual mitigation potential of consumption options for transport measured in tCO2eq/cap. The figure is based on a sample of 23 review articles and 16 consumption options. Negative values (in the red area) represent the potential for backfire.
The dots represent single reviewed studies and the x-s – the average mitigation potential within the same consumption option.
The 25th percentile, median and 75th percentile are noted with lines, with the options ordered by medians. SI Data extraction sheet contains an overview of all options. For transport, we adopted the estimate of 15000 km per passenger per year in the OECD⁷⁰, 1000 km in China⁷¹ and 24000 km in the USA^{71,72} for studies which do not specify annual travel.

The differences in assumed travelled distance explain why options for reducing car travel altogether may show lower mitigation potential compared to a shift to alternatives of internal combustion engine vehicles (ICEV). The Shift to battery electric vehicle (BEV) from ICEV has mitigation potential between 5.4 and -1.9 tCO₂eq/cap, with an average and median of 2.0 tCO₂eq/cap. Carbon reduction potential varies between 3.1 and -0.2 (mean of 0.7) tCO2eq/cap for (plug-in) hybrid electric vehicles (PHEV/HEV), and between 5.8 and – 3.4 (mean of 0) tCO₂eq/cap for fuel cell vehicles (FCV). The carbon intensity of the electricity mix (widely varying across countries⁷⁰) is crucial for the GWP of BEVs^{70,73–76}, where the electricity mix alone was found to explain almost 70% of the variability in LCA results⁷⁶. Furthermore, while modelling studies are often based on the average grid carbon intensity, the marginal emissions factor may be substantially higher if additional demand is met by fossil-fuel thermal plants^{70,76}, e.g. 35% higher in the UK⁷⁰. Fuel consumption is the most influential factor affecting the

1 GWP of ICEV, HEV and PHEV⁷⁴. PHEV have a similar electricity consumption to that of BEV when 2 driving electric⁷⁵. Strong coal-dependence (when the proportion of coal electricity is 20% or larger)⁷⁷ 3 eliminates any potential GHG savings with the shift to FCV. The main advantage of a FCV compared 4 to a BEV is the higher range and quick refilling of the tank^{75,77}; yet, the necessary H₂ filling station 5 infrastructure is currently lacking⁷⁵. We noted substantial differences in the system boundary and 6 modelling approaches, which may also influence the mitigation ranges.

Energy and material efficiency (e.g. more efficient combustion engine, lightweight materials, improved fuel economy, cleaner fuels)^{73,78–81} brings a reduction between 1.46 and 0.01 (mean of 0.3) tCO₂eq/cap. Yet, there has been a clear trend of increased number of vehicles⁶⁴, travelled distance per person⁷⁰ and increased mass of light-duty vehicles⁸⁰, which offset efficiency improvements with transport emissions still on the rise⁶⁴. Differences in ranges may be explained by assumptions about recycling rates and material substitution factors, vehicle lifetime, class and drive cycle and other factors^{78,80}.

We could not evaluate annual mitigation potential from biofuels, as most studies communicate mitigation potential in terms of functional unit (e.g. per MJ of fuel), without further discussions of travelled distance and vehicle efficiency. There are large uncertainties around the mitigation potentials of biofuels due to inconsistencies in scope definition (e.g. system boundary and functional unit), assumptions (e.g. impacts of infrastructure and coproduction), technological choices, and data sources⁸². If system boundaries are expanded to include indirect LUC, physical land constraints from food and feed, and biodiversity conservation as well as the temporal effects on natural carbon stocks, biofuels are revealed as less attractive if not detrimental option for climate change mitigation^{83,84}.

²⁸ 21 Food

Figure 4 provides an overview of various consumption options in the food domain. The majority of
 reviewed studies covered the potential GHG reduction associated with a change of diet and a reduction
 in food waste.

The mitigation potential associated with a diet change involving a reduction in the amount of animal products consumed varies between 2.1 and 0.4 tCO₂eq/cap (mean of 0.9 tCO₂eq/cap) for a Vegan diet, between 1.5 and 0.01 (0.5) for a Vegetarian diet, and between 2.0 and -0.1 (0.6) for Mediterranean and similar diet – e.g. Atlantic and New Nordic. The three types of diets have median mitigation potential of 0.9, 0.5 and 0.4 tCO₂eq/cap, respectively. Adopting more Sustainable diet or a Shift to lower carbon *meats* is also associated with sizable reductions, with an average annual reduction of $0.5 \text{ tCO}_2 \text{eq/cap}$. The carbon intensity per calorie/kg of primary product is substantially lower for vegetal foods compared to ruminants, non-ruminants and dairy^{11,85–87}, with meat producing more emissions per unit of energy due to energy losses at each trophic level⁸⁸. Emissions associated with land use change (LUC) are also most significant for meat-intensive diets⁸⁹, due to increases in pasture land and arable land for growing feed. Nutrition guidelines diets optimized with regards to health guidelines (generally including a reduction in the red meat intake and increase in plant-based foods) are associated with more moderate potential reductions between 1.3 and 0.01 tCO₂eq/cap (mean of 0.3 tCO₂eq/cap).

Improved cooking equipment is associated with strong mitigation potential amounting to a mean and a median of 0.6 tCO₂eq/cap. Cooking methods, fuels, choice of food and cook-ware, use and management of the cook-ware as well as storage time and space are all relevant factors^{90,91}.

Other options for carbon footprint reductions in the food domain focus on the production methods, transportation, seasonality and processing of food products. Organic food have lower emissions compared to conventionally produced food, with an average annual mitigation potential of 0.5 tCO_2eq/cap and a median of 0.4 tCO_2eq/cap . This mitigation potentials is primarily attributable to the increased soil carbon storage and reductions of fertilizers and other agro-chemicals⁹²⁻⁹⁴. Yet, increases in GHG emissions from organic food for the same diet are not uncommon^{92,93,95}, due to lower crop and livestock yields of organic agriculture and the potential increase in production and associated LUC⁹². Opting for Regional and local food and Seasonal and fresh food involves average reductions of 0.4 and

0.2 tCO₂eq/cap. One of the advantages of producing and consuming food in its natural season is that it does not require high-energy input from artificial heating or lighting^{91,96}, thus reducing the embodied GHG emissions. Producing and consuming locally may reduce emissions from transportation and abate impact displacement overall⁹¹, provided there are not large increases in energy requirements (e.g. in the case of heated greenhouse production or through the use of fertilizer^{97,98}). Regional production requiring the use of heating systems (e.g. fresh vegetables in the beginning of the growing season) may be associated with higher emissions compared to even substantial long-distance transport emissions from

1 8 production sites without heating 99 .

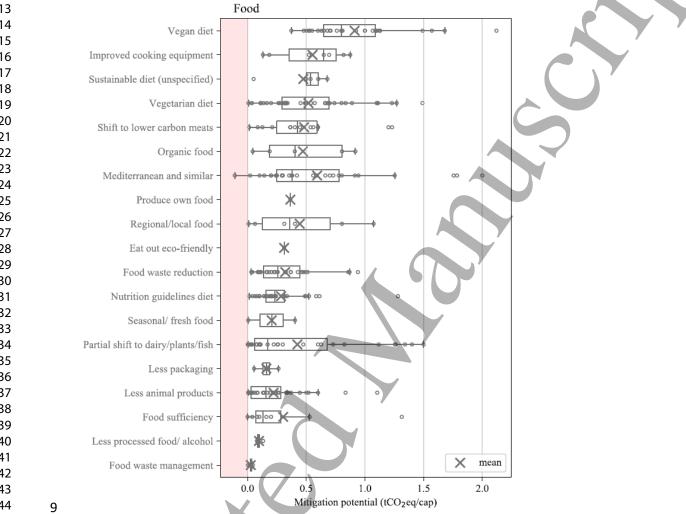


Figure 4: Annual mitigation potential of consumption options for food measured in tCO2eq/cap. The figure is based on a sample
 of 32 review articles and 19 consumption options. Negative values (in the red area) represent the potential for backfire. The
 dots represent single reviewed studies and the x-s – the average mitigation potential within the same consumption option. The
 25th percentile, median and 75th percentile are noted with lines, with the options ordered by medians. SI Data extraction sheet
 contains an overview of all options.

We also note substantial mitigation potential associated with the reduction in consumed food and waste. *Food sufficiency* – implying a reduction in the overall food intake – and *Food waste reduction* options mitigate an average of 0.3 tCO₂eq/cap and a median of 0.1 tCO₂eq/cap. Food waste studies generally make a distinction between avoidable and potentially avoidable waste, which are said to amount to 80%¹⁰⁰ of all food waste. *Food waste management* of unavoidable food waste is associated with more modest average mitigation potential of 0.03 tCO₂eq/cap.

There are large uncertainties^{92,101-104} associated with environmental (e.g. emissions arising from biological processes, LUC and highly integrated production such as beef and dairy), nutritional data (e.g. consumption and waste, weighting factors for gender and age). Impact assessment studies generally

- do not consider emissions associated with LUC¹⁰¹, which is estimated to contribute between 9 and 33%
 of the total livestock emissions (primarily attributable to feed imports)^{92,101}. Furthermore, even though
 food is a basic good (see SM figure 2), the distribution of diets and their embodied GHG impacts is
- 4 largely unequal¹⁰⁵. For example, 20% of diets with the highest carbon contribution in the USA account
 - 5 for more than 45% of the total food-related emissions, mostly linked to meat consumption¹⁰⁵.

6 Housing

- 7 The methodological differences were particularly strong for the reviewed studies in the housing domain,
- 8 where mitigation potential was quantified per kWh of energy use, kg of primary material¹⁰⁶, embodied
- 9 and operational energy per m² of living space, unit of fuel, thermal insulation per surface unit¹⁰⁷ and
 10 others.
- 11 The mitigation options with the highest potential on average include purchasing *Renewable electricity* 12 and *Producing own renewable electricity* with average values of 1.5 (ranging between 2.5 and 0.3) and 13 1.3 (ranging between 4.8 and 0.1) tCO₂eq/cap (Figure 5). The two options have median mitigation 14 potential of 1.6 and 0.6 tCO₂eq/cap, respectively. The mitigation potential of adopting renewable 15 technologies is dependent on the energy source¹⁰⁸ and a wide range of contextual factors¹⁰⁹ – e.g. type 16 of electricity to manufacture renewable technologies, location (affecting the amount of energy that can 17 be produced in the use phase), and the way technologies are used and maintained¹⁰⁹.

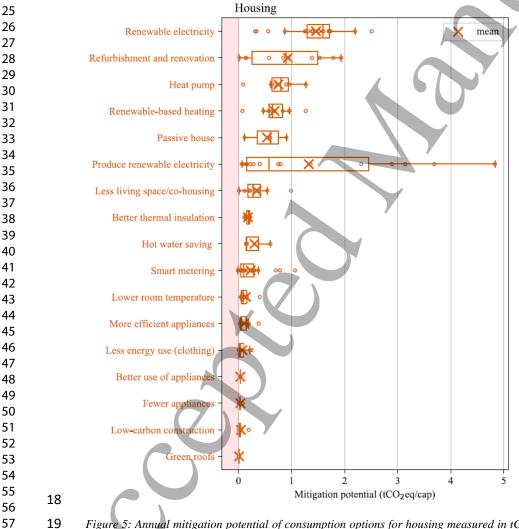


Figure 5: Annual mitigation potential of consumption options for housing measured in tCO₂eq/cap. The figure is based on a sample of 13 review articles and 17 consumption options. Negative values (in the red area) represent the potential for backfire.
The dots represent single reviewed studies and the x-s – the average mitigation potential within the same consumption option (options ordered by averages). The 25th percentile, median and 75th percentile are noted with lines, with the options ordered by medians. SI Data extraction sheet contains an overview of all options.

Other effective infrastructure-related options associated with space heating include Refurbishment and renovation, opting for Heat pump and Renewable-based heating, which offer an average mitigation potential of 0.9, 0.8 and 0.7 tCO₂eq/cap, respectively. The shift to a *Passive house* is associated with an average reduction potential of 0.5 tCO₂eq/cap (based on estimates by three studies), excluding GHG emissions associated with changes in infrastructure. The carbon intensity of materials and sources^{62,108}, infrastructure⁶² and geographical differences in energy and heating requirements and temperature tolerance⁵⁷ are all key factors for the absolute mitigation potential associated with these options. The reviewed mitigation potential of *Smart metering* varies between 1.1 and 0 tCO₂eq/cap, with an average of 0.2 tCO₂eq/cap. Smart metering improves household awareness of their energy consumption and support energy reduction activities (e.g. it may encourage retrofitting of houses or change of appliances and equipment)¹¹⁰. These indirect effects are generally not captured in pilot studies¹¹⁰. Factors such as climate differences, dwelling type and share of renewables in the local grid are of crucial importance for the carbon savings potential¹¹⁰.

Less living space and co-housing – which includes options such as smaller living space (and hence less heating and construction), collective living with others and renting out guest rooms for other people to live in – offer carbon reductions of up to 1.0 tCO₂eq/cap, and an average of 0.3 tCO₂eq/cap. When people live together, they tend to share space heating, cooling, lighting and the structure of the common living space, appliances, tools and equipment^{23,111,112}. While these estimates of household economies of scale from shared living are only limited to the housing domains, sharing within households extends to other types of consumption (e.g. sharing food and cooking together)¹¹². Furthermore, the energy use reductions associated with an additional household member tend to be lower for large households compared to small households¹¹². As building size is the most important factor for home energy consumption, downsizing may substantially reduce housing-related emissions and energy use¹¹³. However, there are significant structural (e.g. lack of adequate alternatives), psychological (e.g. attachment) and security barriers (e.g. loss of ownership) related to downsizing¹¹³. Other behavioral interventions such as Hot water saving and Lowering room temperature by 1-3°C bring about an average saving of 0.3 and 0.1 tCO₂eq/cap, respectively.

³⁵ 28 Other consumption

Finally, other consumption options with substantial mitigation potential include not having a pet and sharing and consumption of services instead of goods with median mitigation potential around 0.3 tCO₂eq/cap (Figure 6). The service/sharing economy includes options such as opting for local, non-market and community services, share and repair. Strategies encouraging sharing include adequate design and infrastructure for durability, recyclability, reuse and product longevity⁸¹ and incentives for multi-household living⁵⁷, grassroots initiatives and downsizing^{32,114}. Yet, studies also warn that peer-to-peer strategies do not necessarily translate into carbon footprint reductions due to extra income and induced consumption¹¹⁵.

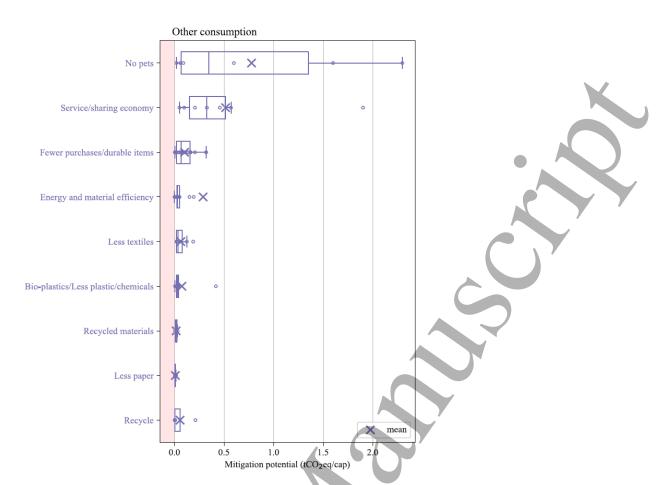


Figure 6: Annual mitigation potential of consumption options for other consumption measured in tCO₂eq/cap. The figure is based on a sample of 10 review articles and 9 consumption options. Negative values (in the red area) represent the potential for backfire. The dots represent single reviewed studies and the x-s – the average mitigation potential within the same consumption option (options ordered by averages). The 25th percentile, median and 75th percentile are noted with lines, with the options ordered by medians. SI Data extraction sheet contains an overview of all options.

7 Discussion and conclusions

8 Mitigation potential of consumption options

One contribution of this study is the systematic provision of mitigation ranges across various consumption domains and the harmonization of results from different methodologies, scopes and assumptions within the same framework (Figure 7). The top consumption options (by medians) include substantial changes in car travel (living car-free, shifting to electric vehicles and public transport), air travel reductions, use of renewable electricity and more sustainable heating (renewable-based heating and heat pump), refurbishment and renovation, a shift to a plant-based diet and improved cooking equipment. The top 10 consumption options together (accounting for the overlap of car travel alternatives) yield an average annual mitigation potential of 9.2 tCO₂eq/cap. While crudely estimated, this indicates a substantial mitigation potential of already available low-carbon consumption options towards achieving the 1.5-2°C target.

Across world regions, the average consumption-based carbon footprints vary between 1.9 and 0.4 tCO₂eq/cap for food, 4.6 and 0.2 tCO₂eq/cap for transport, 3.7 and 0.5 tCO₂eq/cap for housing, and 3.16 and 0.4 tCO₂eq/cap for other consumption^{2,3} (see SM figure 1). United States and Australia stand out with the highest average per capita carbon footprints in our model: with 2.2 and 2.5 tCO₂eq/cap for food, 4.7 and 5.5 tCO₂eq/cap for transport, 5.8 and 4.3 tCO₂eq/cap for housing, and 4.0 and 3.9 tCO₂eq/cap for other consumption, respectively. Yet, the carbon allowances according to the climate targets by 2050

are substantially lower: $0.4 \text{ tCO}_2\text{eq/cap}$ to food, 0.2 to shelter, 0.7 to travel, 0.4 to goods and 0.4 to services, amounting to a total of $2.1 \text{ tCO}_2\text{eq/cap}^{11}$.

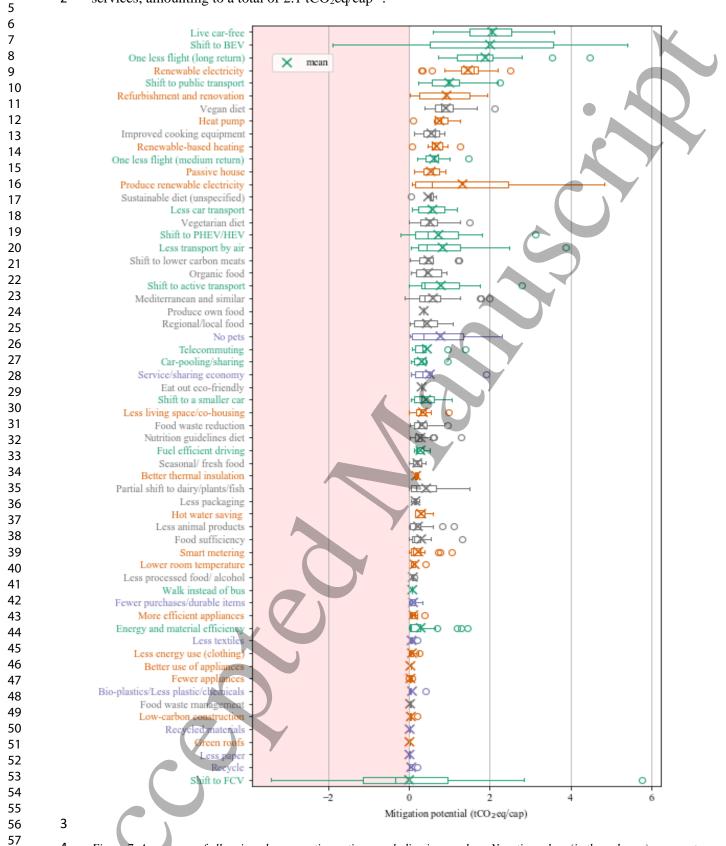


Figure 7: A summary of all reviewed consumption options, excluding inner values. Negative values (in the red area) represent
the potential for backfire. The x-s represent the average mitigation potential within the same consumption option (options
ordered by medians). SI Data extraction sheet contains an overview of all options.

1 The interconnected nature of these strategies need to be recognized in order to adequately respond in 2 mitigating climate change. For example, studies warn about the potential increase in LUC-emissions 3 with the shift to organic; yet, if this shift occurs in parallel to shifts in diets and better food waste 4 management, the conversion of natural or semi-natural vegetation to cropland may be reduced 5 substantially (Figure 4). Furthermore, co-benefits associated with upscaling these mitigation options 6 have also been widely discussed^{64,67,116}.

¹¹ 7 Critical appraisal and limitations

This review is limited to the English language literature published since 2011. More relevant evidence could be captured if the scope is extended to other languages, e.g. capturing more evidence from non-OECD countries. Moreover, although we used a very comprehensive set of search terms, there is a risk that we missed literature that did not list them in their title, abstract or key words. Furthermore, as we did not perform an extensive search for the other consumption domain, we may have omitted key options and potentials. We may have also missed relevant research through the adoption of machine learning and the focus on peer-reviewed literature. Including grey literature (such as theses and governmental reports) would decrease susceptibility to publication bias and resulting inclination of peer-reviews literature towards more 'positive' results'.

The included studies often do not report sufficient methodological details in order to judge rigor of the primary data included. The studies differ largely in assessment method and methodological choices, system boundary, and modelling assumptions. For example, most food-related LCAs adopt a system boundary at the farm gate or retail gate¹⁰¹ (thus, suffering from truncation errors), and exclude consumer losses, impacts associated with the consumption and end-of-life stages, and LUC. LCA reviews generally do not publish an a-priori protocol, conduct a comprehensive and transparent search for studies or discuss an explicit set of inclusion/exclusion criteria. Studies lack transparency with regards to critical appraisal and data extraction, and rarely evaluate the heterogeneity statistically (see the CEESAT tool¹¹⁷). IO studies generally disregard end-of-life stages, LUC emissions and the effects on natural carbon stocks.

Most studies do not consider feedback effects in the global supply chains (e.g. the wider adoption of vegetarian diets is expected to influence the supply chains of hotels, restaurants, supermarkets). Furthermore, the reviewed studies generally disregard embodied emissions in the new infrastructure needed for the upscale of low-carbon practices, e.g. the infrastructure of renewables, and the associated costs. Large-scale investments in energy-intensive industries and infrastructure have been shown to counter-balance and even outweigh the sectoral carbon efficiency gains, especially in fast-developing countries^{118,119}. Prior analysis of GHG emissions from existing and proposed infrastructure suggests that a cost-effective strategy to reduce committed emissions is to target the early retirement of electricity and industry infrastructure in the presence of affordable low-carbon alternatives¹²⁰. Finally, other environmental indicators such as resource use and scarcity may differ substantially in their implications and prioritization of consumption options^{54,98}.

A major obstacle with regards to external validity (applicability to our research question) is that LCA reviews, in particular, communicate mitigation potential by various functional units¹⁰² without providing the context of scale. We particularly excluded a number of housing-related LCA reviews as mitigation potential is solely communicated in terms of functional units. This makes the comparison with other environmental assessments (using different methodologies) and carbon targets/budgets very difficult.

55 43 Modifier effects

Considerations about default-option are critical for the assessment of mitigation potential. While some studies present mitigation potential compared to averages, others compare to "high carbon" consumption patterns¹²¹. Furthermore, there is a large uncertainty associated with basic assumptions about human behavior and public acceptability of demand-side mitigation options⁸¹. While we depict absolute reduction potential of various mitigation options – e.g. shift all car travel to public transport – partial

- 1 adoptions may also be adopted, with relative reduction potential easily calculated as a proportion of the
- 2 ranges discussed in this paper.
- 3 Geographical context and other location, impact assessment method, energy mix and carbon intensity
- 4 and socio-demographics specifications were evaluated in the fixed-effects model as potential factors
- 5 that influence the mitigation potential ranges within consumption options (Figure 8).

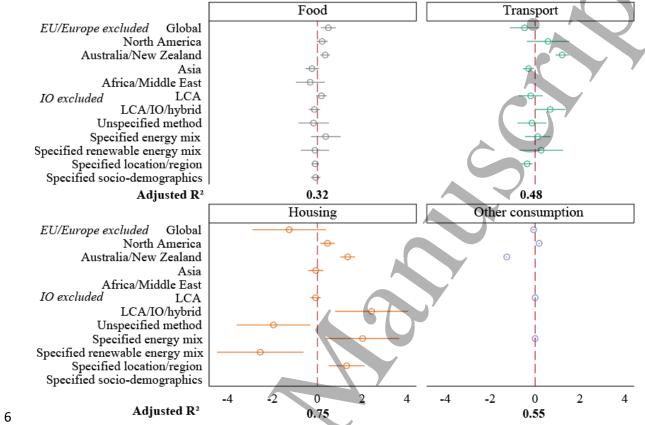


Figure 8: Factors contributing to differences in mitigation potential within consumption options. The coefficients are based on fixed-effects linear model using clustered standard errors by mitigation options. The dependent variable is the annual carbon mitigation potential in tCO2eq per capita.

For food, mitigation potential estimates from North America, Australia and New Zealand of dietary changes are higher compared to EU estimates, while estimates from Asia are lower. In the context of food, LCA-based results were slightly higher compared to IO-based results. Methodological, geographic and socio-demographic factors explain 32% of the differences in mitigation potential within food-related options. Other potential modifier effects include the accounting of food and cooking losses¹²² and LUC¹⁰¹; the magnitude of change/reduction of calories⁸⁹ and the share imported by air¹¹; nutritional guidelines¹²³; the consideration of rebound effects and knock-on savings from food waste reduction or dietary change including avoided shopping and storage¹⁰⁰; and other social and behavioral characteristics^{102–104}.

For transport, mitigation potential estimates from Australia and New Zealand are significantly higher than the European ones, while those from Asia are lower. IO-based estimates are substantially lower than the hybrid estimates in reviewed studies. Geographic and methodological factors attribute to 48% of the differences in mitigation potential within transport-related options. Additional modifiers include fuel and vehicle characteristics^{57,73-75}, travel distance and occupancy rate^{57,61}, energy chain and infrastructure¹²⁴, driving¹²⁴, income group⁶¹ as well as additional technical and behavioral factors¹²⁴.

25 The geographic location, methodology and energy mix are significant for the mitigation potential ranges
26 within housing options, attributing to 75% of the within options variance (Figure 8). The location factor

includes contextual factors influencing the supply of energy, e.g. the location of solar panels during
 use¹⁰⁹ and geographical differences in energy and heating requirements⁵⁷. Additional modifier effects
 include the backup electricity mix, dwelling size, type and lifetime assumption, and additional social

4 and infrastructural influences.

5 Policy recommendations

Finally, we selected the top ranking consumption options and synthesized respective policy
recommendations from the literature. Table 3 communicates a list with the options with the highest
mitigation potential and potential actions towards overcoming the main infrastructural, institutional and
behavioral carbon lock-ins¹²⁵. While the table is informed by the reviewed literature, it should be noted
that we did not conduct a systematic search specifically on targeting actions towards overcoming carbon
lock-ins.

Consumption options with high mitigation potentials	Overcoming infrastructural lock- in	Overcoming institutional lock-in	Overcoming behavioral lock-in
Dietary shift (e.g. vegan, vegetarian)	Change land use practices; Remove investment infrastructure supporting unsustainable and extractive industries	Remove unsustainable subsidies in agriculture, e.g. for meat and dairy; Offer support for alternatives.; Encourage just transition for animal farmers; Better availability of low-carbon options in supermarkets, restaurants, schools, etc; Coordinated efforts of health organizations and government ⁸⁸ ; Ban advertising of high-carbon meats and other high-carbon items.	Encourage low-carbon shared meals ¹²⁶ and diets; Feedbacks for change in social norms and traditions around food consumption ¹²⁶ , e.g. vegan food as default; Decouple veganism/vegetarianism from a particular social identity
Transport mode shift (e.g. active, public transport), car-free	More public transport infrastructure developments for urban and long- distance travel, e.g. cycling lanes, buses, trains; More bike spaces on public transport	Parking and zoning restrictions, e.g. car- free zones and days; Vehicle and fuel tax increases and toll charges; Make driving less convenient in urban areas; Enforce stricter air pollution standards; Ban car advertising	Raising awareness about co-benefits associated with active travel ⁵⁷ ; Socia feedback with the visibility of cycling ¹²⁶ ; Decouple car travel from particular social identity; Improve drivers awareness of cyclers and safety
Reduction in overall travel demand	More compact urban spaces and diverse land use ¹⁶	Allow for flexible working schemes and telecommuting; Halt air travel expansion; Ban flight advertising	Carpooling and carsharing; Encourage telecommuting, moving into denser settlements
Upscaling of electric vehicles	Decarbonize the grid and meet potential additional capacity through renewables; Provision of charging infrastructure	Sustained policy support, e.g. free public charging, tax and fee deductions, subsidies for low-income buyers; Enforce stricter air pollution standards	Tackle charging time acceptance, range anxiety ^{63,70,74}
Renewable-based heating and electricity	Infrastructure investment in renewables	Halt fossil fuel expansion/use and support upscaling of renewables; Incentivize decentralized electricity generation, particularly for low-income households; Enforce stricter air pollution standards; Encourage just transitions for fossil fuel workers; Fossil fuel divestment	Raise public awareness and target NIMBY concerns
Refurbishment and renovation	Energy efficient construction and equipment	Enforce building standards; Encourage investment by dwelling owners and landlords in the fabric of the building and energy efficiency as well as broader home improvements ¹¹⁴ ; Encourage just transitions, e.g. consideration of fuel poverty; Remove inefficiency of listed building	Public awareness around economic and environmental benefits; Reconci investment incentives with householders' images of home comfort ¹¹⁴

Table 3: A summary of the consumption options with the highest mitigation potent.
 infrastructural, institutional and behavioral carbon lock-ins associated with them.

14 Concluding remarks

15 In times of a climate emergency, research and policy urgently needs to move beyond focusing on the16 efficiency of production and use of goods and services. The explicit consideration of the absolute scale

of consumption and its implications for climate change and well-being is ever more relevant. There is a
 need for an open discussion about the overall scale of resource use and emissions and sustainable
 consumption corridors¹²⁷ towards remaining within planetary boundaries and satisfying human needs³³.

We conducted a comprehensive literature review to summarize and compare the reported GHG ranges of various consumption options, critically appraise results and uncertainties, clarify the methodological issues and modifier effects, and identify knowledge gaps to inform future research and policy. The priorities in terms of consumption options may differ substantially depending on income, geographic location, energy context, other factors and carbon lock-ins. Still, consumption is intimately connected

9 to issues of climate change, well-being and sustainability, and thus needs critical attention.

We find that the large majority of the household carbon footprints can be mitigation with already available low-carbon consumption options. Challenging current patterns of consumption and the societal dynamics through a critical assessment of infrastructural, institutional and behavioral lock-ins and potential rebound effects, therefore, needs to become a priority for successful climate change mitigation.

14 Acknowledgements

D.I. and J.B. received funding from the UK Research Councils under the Centre for Research on Energy Demand
Solutions. D.W. received funding from the European Research Council (ERC) under the European Union's
Horizon 2020 research and innovation programme (MAT_STOCKS, grant agreement No 741950).

18 Data availability

19 The data that support the findings of this study are openly available in the supplementary spreadsheet. Please cite20 this article and its digital object identifier (DOI) when making use of the data.

21 References

- 22 1. Masson-Delmotte, V. et al. IPCC Special report 1.5 Summary for policymakers. (2018).
- 23 2. Wood, R. *et al.* Growth in environmental footprints and environmental impacts embodied in trade: Resource efficiency indicators from EXIOBASE3. *J. Ind. Ecol.* (2018). doi:10.5281/zenodo.1038990
- 5 25 3. Ivanova, D. *et al.* Environmental impact assessment of household consumption. J. Ind. Ecol. 20, 526–536 (2016).
- 4. Ivanova, D. & Wood, R. The unequal distribution of household carbon footprints in Europe and its link to sustainability. *Rev. Glob. Sustain.* (2020).
- 285.Otto, I. M., Kim, K. M., Dubrovsky, N. & Lucht, W. Shift the focus from the super-poor to the super-rich. *Nat. Clim. Chang.* 9, 82–84 (2019).
- 30 6. Hubacek, K., Baiocchi, G., Feng, K. & Patwardhan, A. Poverty eradication in a carbon constrained world. *Nat. Commun.* 8, 1–8 (2017).
- **32** 7. Wiedenhofer, D. *et al.* Unequal household carbon footprints in China. *Nat. Clim. Chang.* **7**, (2016).
- 8. Hubacek, K. *et al*, Global carbon inequality. *Energy, Ecol. Environ.* **2**, 361–369 (2017).
- 8349.United Nations, World Population Prospects: The 2015 Revision, Methodology of the United Nations Population9355Estimates and Projections. United Nations Economic and Social Affairs, Population Division XXXIII, (2015).
- 3610.Tukker, A. *et al.* Environmental and resource footprints in a global context: Europe's structural deficit in resource
endowments. *Glob. Environ. Chang.* 40, 171–181 (2016).
- 38 38 39 11. Girod, B., van Vuuren, D. P. & Hertwich, E. G. Climate policy through changing consumption choices: Options and obstacles for reducing greenhouse gas emissions. *Glob. Environ. Chang.* 25, 5–15 (2014).
- 40 12. O'Neill, D. W., Fanning, A. L., Lamb, W. F. & Steinberger, J. K. A good life for all within planetary boundaries. Nat. Sustain. 1, 88–95 (2018).
- 42 13. Anderson, K. Duality in climate science. *Nat. Geosci.* 8, 1–2 (2015).
- 43 14. Gasser, T., Guivarch, C., Tachiiri, K., Jones, C. D. & Ciais, P. Negative emissions physically needed to keep global
 44 warming below 2°C. *Nat. Commun.* 6, (2015).

AUTHOR SUBMITTED MANUSCRIPT - ERL-108137.R1

1 2			
3	1	15.	Anderson, K. & Peters, G. The trouble with negative emissions. Science (80). 354, 182–183 (2016).
4 5 6	2 3	16.	Creutzig, F. <i>et al.</i> Beyond Technology: Demand-Side Solutions for Climate Change Mitigation. <i>Annu. Rev. Environ. Resour.</i> 41 , 173–198 (2016).
7	4	17.	Minx, J. C. et al. Negative emissions - Part 1: Research landscape and synthesis. Environ. Res. Lett. 13, (2018).
8 9 10	5 6 7	18.	IPCC. Summary for Policymakers. Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change (2014). doi:10.1017/CBO9781107415324
11 12 13	8 9	19.	O'Neill, B. C. <i>et al.</i> The roads ahead: Narratives for shared socioeconomic pathways describing world futures in the 21st century. <i>Glob. Environ. Chang.</i> 42 , 169–180 (2017).
14 15	10 11	20.	Creutzig, F. <i>et al.</i> Towards demand-side solutions for mitigating climate change. <i>Nat. Clim. Chang.</i> 8 , 260–271 (2018).
16 17 18	12 13	21.	O'Rourke, D. & Lollo, N. Transforming Consumption: From Decoupling, to Behavior Change, to System Changes for Sustainable Consumption. <i>Ssrn</i> (2015). doi:10.1146/annurev-environ-102014-021224
19	14	22.	Ivanova, D. et al. Mapping the carbon footprint of EU regions. Environ. Res. Lett. 12, 1-13 (2017).
20 21 22	15 16 17	23.	Wiedenhofer, D., Smetschka, B., Akenji, L., Jalas, M. & Haberl, H. Household time use, carbon footprints, and urban form: a review of the potential contributions of everyday living to the 1.5 °C climate target. <i>Curr. Opin. Environ. Sustain.</i> 30 , 7–17 (2018).
23 24 25	18 19	24.	Malik, A., Mcbain, D., Wiedmann, T. O., Lenzen, M. & Murray, J. Advancements in Input-Output Models and Indicators for Consumption-Based Accounting. <i>J. Ind. Ecol.</i> 00 , 1–13 (2018).
25 26 27	20 21	25.	Wiedmann, T. & Lenzen, M. Environmental and social footprints of international trade. <i>Nat. Geosci.</i> 11 , 314–321 (2018).
28 29 30	22 23	26.	Wiedmann, T., Wilting, H. C., Lenzen, M., Lutter, S. & Palm, V. Quo Vadis MRIO? Methodological, data and institutional requirements for multi-region input-output analysis. <i>Ecol. Econ.</i> 70 , 1937–1945 (2011).
31 32	24 25	27.	Wilson, J., Tyedmers, P. & Grant, J. Measuring environmental impact at the neighbourhood level. <i>J. Environ. Plan. Manag.</i> 56 , 42–60 (2013).
33 34 25	26 27	28.	Reap, J., Roman, F., Duncan, S. & Bras, B. A survey of unresolved problems in life cycle assessment. <i>Int. J. Life Cycle Assess.</i> 13 , 290–300 (2008).
35 36 37	28 29	29.	Wynes, S. & Nicholas, K. A. The climate mitigation gap: education and government recommendations miss the most effective individual actions. <i>Environ. Res. Lett.</i> 12 , (2017).
38 39	30 31	30.	Vita, G. et al. Deliverable 7.3: Analysis of current impact of lifestyle choices and scenarios for lifestyle choices and green economy developments. GLAMURS: EU SSH.2013.2.1-1 Grant agreement no. 613420. (2016).
40 41 42	32 33 34	31.	Rodrigues, J., Prado, V., Van Der Voet, E., Moran, D. & Wood, R. FP7 Carbon-Cap -Effectiveness of life-cycle micro-level options for consumption based climate mitigation D6.2 Effectiveness of improvement options to reduce GHG emissions in a consumption based approach. EU FP7 Carbon-CAP project 24, (2015).
43 44 45	35 36	32.	Vita, G. <i>et al.</i> Happier with less? Members of European environmental grassroots initiatives reconcile lower carbon footprints with higher life satisfaction and income increases. <i>Energy Res. Soc. Sci.</i> 60 , 101329 (2020).
46 47	37 38	33.	Pirgmaier, E. & Steinberger, J. Roots, Riots, and Radical Change—A Road Less Travelled for Ecological Economics. <i>Sustainability</i> 11 , 2001 (2019).
48 49 50	39 40	34.	Gardner, C. J. & Wordley, C. F. R. Scientists must act on our own warnings to humanity. <i>Nat. Ecol. Evol.</i> 3 , 1271–1272 (2019).
51 52	41 42	35.	Heede, R. Tracing anthropogenic carbon dioxide and methane emissions to fossil fuel and cement producers, 1854-2010. <i>Clim. Change</i> 122 , 229–241 (2014).
53 54 55	43 44	36.	Hertwich, E. G. & Wood, R. The growing importance of scope 3 greenhouse gas emissions from industry. <i>Environ. Res. Lett.</i> 13 , (2018).
55 56	45	37.	Edenhofer, O. King coal and the queen of subsidies. Science (80). 349, 1286–1287 (2015).
57 58	46	38.	Rainforest Action Network et al. Banking on climate change. Frontiers in Ecology and the Environment (2019).
59	47	39.	Creutzig, F. et al. Urban infrastructure choices structure climate solutions. Nat. Clim. Chang. 6, 1054–1056 (2016).
60	48 49	40.	Wood, R. <i>et al.</i> Prioritizing Consumption-Based Carbon Policy Based on the Evaluation of Mitigation Potential Using Input-Output Methods. <i>J. Ind. Ecol.</i> 0 , 1–13 (2017).

3	1	41.	Gillingham, K., Rapson, D. & Wagner, G. The Rebound Effect and Energy Efficiency Policy. 0, 1–22 (2015).
4 5 6	2 3	42.	Lekve Bjelle, E., Steen-Olsen, K. & Wood, R. Climate change mitigation potential of Norwegian households and the rebound effect. <i>J. Clean. Prod.</i> 172 , 208–217 (2018).
7	4	43.	Akenji, L. Consumer scapegoatism and limits to green consumerism. J. Clean. Prod. 63, 13–23 (2014).
8 9 10	5 6	44.	Grubler, A. <i>et al.</i> A low energy demand scenario for meeting the 1.5 °c target and sustainable development goals without negative emission technologies. <i>Nat. Energy</i> 3 , 515–527 (2018).
10 11 12	7 8	45.	Hertwich, E. & Peters, G. Carbon footprint of nations: A global, trade-linked analysis. <i>Environ. Sci. Technol.</i> 43 , 6414–6420 (2009).
13 14	9 10	46.	Owen, A. <i>et al.</i> Energy consumption-based accounts: A comparison of results using different energy extension vectors. <i>Appl. Energy</i> 190 , 464–473 (2017).
15 16 17	11 12	47.	Moran, D. <i>et al.</i> Quantifying the potential for consumer-oriented policy to reduce European and foreign carbon emissions. <i>Clim. Policy</i> 0 , 1–11 (2018).
18 19	13 14	48.	Suh, S. et al. System Boundary Selection in Life-Cycle Inventories Using Hybrid Approaches. Environ. Sci. Technol. 38 , 657–664 (2004).
20 21 22 23	15 16 17	49.	Collaboration for Environmental Evidence. Guidelines and Standards for Evidence synthesis in Environmental Management. <i>Version 5.0</i> (2018). Available at: www.environmentalevidence.org/information-for-authors. (Accessed: 23rd December 2019)
23 24 25 26	18 19 20	50.	Haddaway, N. R., Macura, B., Whaley, P. & Pullin, A. S. ROSES RepOrting standards for Systematic Evidence Syntheses : pro forma , flow - diagram and descriptive summary of the plan and conduct of environmental systematic reviews and systematic maps. <i>Environ. Evid.</i> 7 , 1–8 (2018).
27 28 29 30	21 22 23 24	51.	Ivanova, D., Barrett, J., Wiedenhofer, D., Macura, B. & Creutzig, F. Outline for review of reviews: Quantifying the potential for climate change mitigation of consumption-based options. (2019). Available at: https://docs.google.com/document/d/1Esu4DU1IvVqKR8A54te3jNmvAW_3hiXFfY6h67mAhz0/edit. (Accessed: 10th March 2020)
31 32	25 26	52.	Creutzig, F. <i>et al.</i> Mapping the whole spectrum of demand, services and social aspects of mitigation. <i>Environ. Res. Lett.</i> (2020).
33 34	27	53.	Grieneisen, M. L. & Zhang, M. The current status of climate change research. Nat. Clim. Chang. 1, 72-73 (2011).
35 36	28 29	54.	Vita, G. <i>et al.</i> The environmental impact of green consumption and sufficienct lifestyles scenarios in Europe: Connecting local visions to global consequences. <i>Rev. Ecol. Econ.</i> (2019).
37 38 39	30 31	55.	Abrahamse, W., Steg, L., Vlek, C. & Rothengatter, T. A review of intervention studies aimed at household energy conservation. <i>J. Environ. Psychol.</i> 25 , 273–291 (2005).
40 41	32 33	56.	Dietz, T., Gardner, G. T., Gilligan, J., Stern, P. C. & Vandenbergh, M. P. Household actions can provide a behavioral wedge to rapidly reduce US carbon emissions. <i>Proc. Natl. Acad. Sci. U. S. A.</i> 106 , 18452–18456 (2009).
42 43 44	34 35	57.	Ivanova, D. et al. Carbon mitigation in domains of high consumer lock-in. Glob. Environ. Chang. 52, 117–130 (2018).
44 45 46	36 37	58.	Schanes, K., Giljum, S. & Hertwich, E. Low carbon lifestyles: A framework to structure consumption strategies and options to reduce carbon footprints. <i>J. Clean. Prod.</i> 139 , 1033–1043 (2016).
47 48	38 39	59.	Ottelin, J., Heinonen, J. & Junnila, S. Rebound Effects for Reduced Car Ownership and Driving. in Nordic Experiences of Sustainable Planning: Policy and Practice (ed. Kristjánsdóttir, S.) (Routledge, 2017).
49 50 51	40 41	60.	AIRBUS. 2019-2038 GMF - Data spreadsheet. <i>Global Market Forecast</i> (2019). Available at: https://www.airbus.com/aircraft/market/global-market-forecast.html. (Accessed: 11th March 2020)
52 53	42 43	61.	Lacroix, K. Comparing the relative mitigation potential of individual pro-environmental behaviors. <i>J. Clean. Prod.</i> 195 , 1398–1407 (2018).
54 55 56 57	44 45 46	62.	Institute for Global Environmental Strategies Aalto University and D-mat ltd. <i>1.5-Degree Lifestyles: Targets and options for reducing lifestyle carbon footprints. Technical report.</i> (Institute for Global Environmental Strategies, Hayama, Japan, 2019).
57 58 59	47 48	63.	Rolim, C. C., Gonçalves, G. N., Farias, T. L. & Rodrigues, Ó. Impacts of Electric Vehicle Adoption on Driver Behavior and Environmental Performance. <i>Procedia - Soc. Behav. Sci.</i> 54 , 706–715 (2012).
60	49 50	64.	Kwan, S. C., Tainio, M., Woodcock, J., Sutan, R. & Hashim, J. H. The carbon savings and health co-benefits from the introduction of mass rapid transit system in Greater Kuala Lumpur, Malaysia. <i>J. Transp. Heal.</i> 6 , 187–200
			21

1

V

(2017).

4 5	2	65.	Aamaas, B., Borken-Kleefeld, J. & Peters, G. P. The climate impact of travel behavior: A German case study with
6	3		illustrative mitigation options. <i>Environ. Sci. Policy</i> 33 , 273–282 (2013).
7 8	4	66.	Aamaas, B. & Peters, G. P. The climate impact of Norwegians' travel behavior. <i>Travel Behav. Soc.</i> 6, 10–18 (2017).
9 10	5 6	67.	Woodcock, J., Givoni, M. & Morgan, A. S. Health Impact Modelling of Active Travel Visions for England and Wales Using an Integrated Transport and Health Impact Modelling Tool (ITHIM). <i>PLoS One</i> 8 , (2013).
11 12	7 8	68.	Chester, M. & Horvath, A. Life-cycle energy and emissions inventories for motorcycles, diesel automobiles, school buses, electric buses, Chicago Rail, and New York City Rail. (2009).
13 14	9 10	69.	Union of Concerned Scientists. <i>Ride-Hailing's Climate Risks: Steering a growing industry toward a clean transportation future</i> . (2019).
15 16 17	11 12	70.	Tran, M., Banister, D., Bishop, J. D. K. & McCulloch, M. D. Realizing the electric-vehicle revolution. <i>Nat. Clim. Chang.</i> 2 , 328–333 (2012).
18 19	13 14	71.	Ng, WS. & Schipper, L. China Motorization Trends: Policy Options in a World of Transport Challenges. <i>Grow. Greenh. Prot. Clim. By Putt. Dev. First</i> 49–67 (2005).
20 21 22	15 16	72.	Mahmoudzadeh Andwari, A., Pesiridis, A., Rajoo, S., Martinez-Botas, R. & Esfahanian, V. A review of Battery Electric Vehicle technology and readiness levels. <i>Renew. Sustain. Energy Rev.</i> 78 , 414–430 (2017).
23 24	17 18	73.	Cornell, R. & A. The Environmental Benefits of Electric Vehicles as a Function of Renewable Energy Ryan. 32 ، ص ، 117 (2017).
25 26	19 20	74.	Onat, N. C., Kucukvar, M., Aboushaqrah, N. N. M. & Jabbar, R. How sustainable is electric mobility? A comprehensive sustainability assessment approach for the case of Qatar. <i>Appl. Energy</i> 250 , 461–477 (2019).
27 28 29	21 22	75.	Helmers, E. & Marx, P. Electric cars: technical characteristics and environmental impacts. <i>Environ. Sci. Eur.</i> 24, 1–15 (2012).
30 31	23 24	76.	Marmiroli, B., Messagie, M., Dotelli, G. & Van Mierlo, J. Electricity generation in LCA of electric vehicles: A review. <i>Appl. Sci.</i> 8 , (2018).
32 33 34	25 26	77.	Hao, H., Mu, Z., Liu, Z. & Zhao, F. Abating transport GHG emissions by hydrogen fuel cell vehicles: Chances for the developing world. <i>Front. Energy</i> 12 , 466–480 (2018).
35 36	27 28	78.	Kim, H. C. & Wallington, T. J. Life-cycle energy and greenhouse gas emission benefits of lightweighting in automobiles: Review and harmonization. <i>Environ. Sci. Technol.</i> 47 , 6089–6097 (2013).
37 38	29 30	79.	Speth, R. L. <i>et al.</i> Economic and environmental benefits of higher-octane gasoline. <i>Environ. Sci. Technol.</i> 48 , 6561–6568 (2014).
39 40 41	31 32 33	80.	Luk, J. M., Kim, H. C., De Kleine, R., Wallington, T. J. & MacLean, H. L. Review of the Fuel Saving, Life Cycle GHG Emission, and Ownership Cost Impacts of Lightweighting Vehicles with Different Powertrains. <i>Environ. Sci. Technol.</i> 51 , 8215–8228 (2017).
42 43 44	34 35	81.	Cherry, C., Scott, K., Barrett, J. & Pidgeon, N. Public acceptance of resource-efficiency strategies to mitigate climate change. <i>Nat. Clim. Chang.</i> 8 , 1007–1012 (2018).
45 46	36 37	82.	Farrell, A. E. <i>et al.</i> Ethanol Can Contribute to Energy and Environmental Goals. <i>Science (80).</i> 311 , 506–509 (2006).
47 48 49	38 39	83.	Creutzig, F. <i>et al.</i> Reconciling top-down and bottom-up modelling on future bioenergy deployment. <i>Nat. Clim. Chang.</i> 2 , 320–327 (2012).
50 51	40 41	84.	Kalt, G. <i>et al.</i> Greenhouse gas implications of mobilizing agricultural biomass for energy: A re-assessment of global potentials in 2050 under different food-system pathways. <i>Environ. Res. Lett.</i>
52 53	42 43	85.	Poore, J. & Nemecek, T. Reducing food 's environmental impacts through producers and consumers. 992 , 987–992 (2018).
54 55 56	44 45	86.	González, A. D., Frostell, B. & Carlsson-Kanyama, A. Protein efficiency per unit energy and per unit greenhouse gas emissions: Potential contribution of diet choices to climate change mitigation. <i>Food Policy</i> 36 , 562–570 (2011).
57 58	46 47	87.	Clark, M. A., Springmann, M., Hill, J. & Tilman, D. Multiple health and environmental impacts of foods. <i>Proc. Natl. Acad. Sci. U. S. A.</i> 1–6 (2019). doi:10.1073/pnas.1906908116
59 60	48	88.	Godfray, H. C. J. et al. Meat consumption, health, and the environment. Science (80). 33, 935–937 (2018).
00	49	89.	Brunelle, T., Coat, M. & Viguié, V. Demand-side mitigation options of the agricultural sector: Potential, barriers and 22
			22

 ways forward. <i>OCL. Oblaced Stat. Com. Lipids</i> 24, (2017). Xu, Z., Sun, D. W., Zhang, Z. & Zhu, Z. Research developments in methods to reduce carbon footprint of cooking operations. <i>Arview. Trends Food Sci. Technol.</i> 44, e0-57 (2015). Xu, Z. <i>et al.</i> Research Developments in Methods in Relate the Carbon Footprint of the Food System: A Review. <i>Crit. Rev. Tood Sci. Mer.</i> 55, 1270–1286 (2015). Xu, Z. <i>et al.</i> Research Developments in Methods in Relate the Carbon Footprint of the Food System: A Review. <i>Crit. Rev. Tood Sci. Mer.</i> 55, 1270–1286 (2015). Lynch. D. H., MacRea, R. & Marin, R. C. The carbon and global warming potential impacts of organic families. Does it have a significant role in an energy constrained world? <i>Statistical Mill</i> 5, 322–362 (2011). Lacou, C. <i>et al.</i> Bavionement Impacts of Plans Hased Dest: How Dees Organic Food Consumption Compilate to Environmental Statisticity <i>Proc. Natr. Sci.</i> 7, 369–375 (2014). Meier, M. <i>et al.</i> Environmental impacts of organic and conventional agricultural produces: - Are the differences captured by the cycle assessment? <i>J. Environ. Markage</i>, 129, 132–342 (2015). Meier, M. <i>et al.</i> Environmental impacts of organic and conventional agricultural produces: - Are the differences captured by the cycle assessment? <i>J. Environ. Markage</i>, 359–375 (2014). Meisr, M. J., Logez, L. A. <i>Culano, M. A.</i> (Gineer, N. Cucarry, T. Hessenshi Howedols' Consumption Conversion and an intermet of the transformation of the UK have a generate environmental subantiality? <i>Proc. Mirk. Sci.</i> 7, 359–375 (2014). Webb, J., Williams, A. G., Hope, F., Evans, D. & Moothouse, F. Do foods innortal into the UK have a generate environmental subantiality? <i>Proc. Mirk. Sci.</i> 7, 359–375 (2014). Tobura, M. A. López, L. A. <i>Culano, M. A.</i> (Gineer, N. Cucarry, T. Hessenshi Howedols' Consumption Good for the Nexus Curbon. Water Moothouse, F. Do foods innortal in alt	2			
 Nu, Z., Sun, D. W., Zhang, Z. & Zhu, Z. Research developments in methods to reduce carbon footprint of cooking operations. <i>A review. Trends Food Sci. Technol.</i> 44, 49–57 (2015). Nu, T., <i>et al.</i> Research Developments in Methods to Reduce the Carbon Footprint of the Food System : A Reduce Critic Rev. Food Sci. Natr. 55, 1270–1286 (2015). Smith, L. G., Kirk, G. J. D., Janes, P. J. & Williams, A. G. The greenhouse gas impacts of converting food production in England and Wales to organic methods. <i>Nat. Commun.</i> 10, 4641 (2019). Lyoch, D. H., Mackea, R. K. Martin, R. C. The carbon and plothal worming potential impacts of organic familine Does it have a significant role in an energy constrained world? <i>Socialability</i> 3, 322–362 (2011). Lacour, C. <i>et al.</i> Environmental Impacts of Plane Based Diets: How Does Organic Food Consumption Coltribute to Havirannenial Bustainability? <i>Pron. Natr.</i> 5, 1–18 (2018). Medier, M. S. <i>et al.</i> Environmental impacts of organic and conventional agricultural products: Are the differences captured by life cycle assessment? <i>J. Environ. Manage</i>, 149, 193 208 (2015). Macharmid, J. J. Seasonality and dietary requirements: will cating seasonal food columbride to break an environmental sustainability? <i>Proc. Natr.</i> 5, 73, 368 375 (2014). Webb, J., Williams, A. G. Hoper, E. Kyuns, D. & Moorhouse: E. Do foods impreted into the UK have a greater environmental impact than the same foods produced within the UK? <i>Int. J. Eng. Cycl. Assess.</i> 8, 1325–1343 (2013). Thour, M. A., <i>Loper,</i> L. A., Cadaron, M. A., Giomer, N. & Cazaron, J. N. Basconal Household's: Consumption Good for the Neuros Carbon Water Footprint' in the Spanis Print and Vecetable Technol. 52, 12066 12077 (2018). Thour, M. C., Haber, H. Erb, K. H. & Lindenthal. T. Contrasted greanhouse gas emissions from local versus long-range tomato production. <i>Agen. Sci. Mat. Page.</i> 8, 24, 24020 (2017). Hal		1		ways forward. OCL - Oilseeds fats, Crop. Lipids 24, (2017).
 M. Z. <i>et al.</i> Reserved Developments in Methods to Relate the Carbon Footprint of the Food System: A Review. <i>Crit. Rev. Food Sci. Natr.</i> 55, 1270–1286 (2015). Smith, L. G., Kirk, G. J. D. Jones, P. J. & Williams, A. G. The greenhouse gas impacts of converting food production in England and Wales to organic methods. <i>Nat. Commun.</i> 10, 4611 (2019). Juych, D. H., Mackas, R. & Marin, R. C. The carbon and plath warming pretential impacts of organic familing: Does it have a significant role in an energy constrained world? <i>Sustainability</i> 3, 322–362 (2011). H., Mackas, R. & Marin, R. C. The carbon and plath warming pretential impacts of organic family Does it have a significant role in an energy constrained world? <i>Sustainability</i> 3, 322–362 (2011). H. Laccur, C. <i>et al.</i> Environmental impacts of organic and conventional agricultural products - Are the differences englamed by the cycle assessment? <i>J. Environ. Manage.</i> 149, 103–208 (2015). Meier, M. S. <i>et al.</i> Environmental impacts of <i>organic personal</i> food onturbate to bealth and environmental sustainability? <i>Proc. Natr.</i> 50: 73, 368–375 (2014). Mebb, J., Williams, A. G., Hope, E., Evas, D. & Moorhouse, E. Do foods imported into the Uk was a greater environmental impact than the same foods produced within the UK? <i>Int. J. Clif. Col. Assess.</i> 18, 1325–1343 (2013). Tobarra, M. A. López, L. A., Cadarco, M. A., Gómez, N. & Cazaron, U. Sastonal Household? Consumption Good for the Neuro Carbon Water Footprint? the Spanish Fruits and Vegetables Care. <i>Environ. Sci. Technol.</i> 52, 12066–12077 (2018). Theuri, M. C., Haberl, H., Erb, K. H. & Lindenthal. T. Contrasted greenhouse gas emissions from local versus long- range to mato production. <i>Agric Stastian Dev. & 45</i>, 453 (02104). Salendech, R., Font Vivanco, D., Al-Tarbaa, A. & <i>environmental langes</i> of hit greenhouse gas emissions from local versus long- range to mato product	5		90.	
 Samin E. O., Ruk, G. J. E., Jong M. G. J. E., K. H. K. S. J. K. G. Hing greatings gas inputs for organic farming: proteining proteining inputs in organic farming: Disciplication of an an energy constrained world? Stationability 3, 322–362 (2011). J. Lacour, C. <i>et al.</i> Environmental Impacts of Plane Based Dises: How Does Organic Food Consumption Complute to Environmental Statianability? <i>Front. Nur. S.</i> 1–13 (2018). Meier, M. S. <i>et al.</i> Environmental Impacts of organic and conventional agricultural products - Are the differences captured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193–208 (2015). Meier, M. S. <i>et al.</i> Environmental impacts of an environmental statianability? <i>Proc. Nur. S.</i> 3, 368 375 (2014). Macdiamid, J. L. Seasonality and distary requirements: will eating seasonal food calucitohie in health and environmental statianability? <i>Proc. Nur. Soc.</i> 3, 368 375 (2014). Webb, J., Williams, A. G., Hope, E., Evans, D. & Moorhouse, E. Do foods imported into the UK have a greater environmental impact than the same foods produced within the UK? <i>Int. J. Life Corle Assess. Bl</i>, 1325 1343 (2013). Toherra, M. A., Lioper, J. A., Cadarso, M. A. Goinez, N. & Cazaero, 1-to Saeonal Household: Comamptium Grood for the Nexus Carbon/Water Foorphit? the Spanish truits and Vagenable Use. Environ. Sci. Technol. Sci. Technol. Sci. 12065 12077 (2018). Theurd, M. C., Habert, H., Eth, K. H. & Lindenthal, T. Contrasted preachouse gas emissions from local versus long-trange tomato production. <i>Ageno. Statiatin. Dev.</i> 34, 593–602 (2014). Salemdach, R., Font Vivanco, D., Al-Tabbaa, A. & en Emrgosceh, E. K. H. J. A holistic approach to the environmental evaluation of food waste provention. <i>Water Models</i>, 59, 442–450 (2017). Bellindy, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3-118 (2013). Hablattom, E., Carlsson-	7		91.	
 S. Lynch, D. H., MacRae, R. & Marin, R. C. The carbon and global warming potential impacts of organic farming: Does it have a significant role in an energy constrained world's <i>Stastinability</i> 3, 522–562 (2011). J. Lacour, C. <i>et al.</i> Environmental Impacts of Plant-Based Diets; How Does Organic Food Consumption Compliante to Environmental Sustainability? <i>Front. Nat.</i> 5, 1 13 (2018). Meier, M. S. <i>et al.</i> Environmental impacts of organic and conventional agricultural products - Are the differences captured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193–208 (2015). Meier, M. S. <i>et al.</i> Environmental impacts of organic and conventional agricultural products - Are the differences captured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193–208 (2015). Media, M. J., & <i>et al.</i> Environmental impacts of complex on the conventional agricultural products - Are the differences captured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193–208 (2015). Webb, J., Williams, A. G., Hope, E., Evans, D. & Moorhouse, E. Do foods imported into the UK have a greater environmental sustainability? <i>Proc. Nat. Sci.</i> 7, 308–375 (2014). Toberra, M. A., López, L. A., Cadaros, M. A., Gómez, M. A. Cazearo, T. N. Saesson Households' Consumption Gread for the Nexus Carbon/Water Footprint? the Spanish Fruits and Vagetables Cass: <i>Environ. Sci. Technol.</i> 52, 12066–12077 (2018). Theurl, M. C., Haberl, H., Erh, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versals long- range tomato production. <i>Ageon. Statistic. Dev.</i> 34, 593 500 (2014). Salemdeeb, R., Fout Vivanco, D., Al-Tabbea, A. & 2n Emigasson, E. K. H. J. A holistic approach to the environmental evaluation of food waste previous. <i>Nature Manage</i>, 59, 442–450 (2017). Hellarby, J. <i>et al.</i> Livestock greenhouse gas emissions for different fresh food categories. <i>J. Cean., Front.</i> 91, 1–11 (2015). Hellarby,	10		92.	
 94. Lacour, C. <i>et al.</i> Environmental Impacts of Plant-Based Diets: How Does Organic Food Consumption Complute to Environmental Sustainability? <i>Proc. Nutr.</i> 5, 1 13 (2018). 95. Meier, M. S. <i>et al.</i> Environmental impacts of organic and conventional agricultural products - Are the differences expured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193-208 (2015). 96. Maediarnid, J. J. Seasonality and dietary requirements: will eating seasonal food contribute to bealth and environmental impact than the same foods produced within the UK? <i>Int. J. Efe Cycle Assess.</i> 18, 1325–1343 (2013). 97. Webb, J., Williams, A. G., Hope, E., Evans, D. & Moorhouse, E. Do foods imported into the UK have a greater environmental impact than the same foods produced within the UK? <i>Int. J. Efe Cycle Assess.</i> 18, 1325–1343 (2013). 98. Tobarra, M. A., López, L. A., Cadarso, M. A., Gómez, N. & Cazearro, J. 'Is Seasonal Howelods' Consumption Good for the Nexus Carbon Water Footprint? the Spanish Fruits and Vegetables Case. Environ. Sci. Technol. 52, 12066–12077 (2018). 99. Theuri, M. C., Haberl, H., Erh, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long-range tomato production. <i>Agron. Stastain. Dev.</i> 34, 593–602 (2014). 21 100. Salemdech, R., Font Vivanco, D., Al-Tabbaa, A. & zu Ermagesch, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Water Manage</i> 39, 442–450 (2017). 22 101. Bellarky, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3–18 (2013). 23 103. Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 140, 166–793 (2017). 33 104. Song, G., Li, M., Fullanz, PJane, P. Environmental impact of dietary change: a systematic review of <i>Licean. Prod.</i> 140, 166–793 (2017). 34	12	8 9	93.	
 Meier, M. S. <i>et al.</i> Environmental impacts of organic and conventional agricultural products: Are the differences captured by life cycle assessment? <i>J. Environ. Manage.</i> 149, 193–208 (2015). Macdiarmid, J. I. Seasonality and dietary requirements: will eating seasonal food contribute to health and environmental sustainability? <i>Proc. Natr. Soc.</i> 73, 368–375 (2014). Webb, J., Williams, A. G., Hope, E., Evans, D. & Moorhouse, E. Do foods imported into the UK have a greater environmental impact than the same foods produced within the UK? <i>Int. J. Circ Cycle Assess.</i> 18, 1325–1343 (2013). Tobarri, M. A., López, L. A., Cadarso, M. A., Gómez, N. & Cazcarro, Irts Seasonal Households' Consumption Good for the Nexa: Carbon/Water Foorprint? the Spanish Fruits and Vegetables Case. <i>Environ. Sci. Technol.</i> 52, 12066–12077 (2018). Theurt, M. C., Habert, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long-range tomato production. <i>Agron. Sustain. Dev.</i> 34, 593–602 (2014). Salenndeeb, R., Fortt Wixarco, D., Al-Tabbaa, A. & zu Ermgasseh, F. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Waste Manage</i>, 59, 442–450 (2017). Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3 18 (2013). Chane, S., Crossin, F. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prool.</i> 101, 0673 (2017). Hallström, E., Carlsson-Kanyama, A. & Borjesson, P. Environmental impact of dietary change: A systematic review. <i>J. Clean. Prool.</i> 101, 076–33 (2017). Chane, S., Crossin, F. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prool.</i> 101, 076–33 (2017). Heller, M. C., Willits-Smith, A., Myyer, R., Koeleian, G. A. & Rose, D. Greenh	14		94.	
 94 96. Macdiarmid, J. J. Seasonality and dietary requirements. will eating seasonal food cohardbate to health and environmental sustainability? <i>Proc. Natr. Soc.</i> 73, 368–375 (2014). 16 97. Webb, J., Williams, A. G., Hope, E., Evans, D. & Moorhouse, E. Do Foods imported into the UK have a greater environmental impact than the same foods produced within the UK? <i>Int. J. Life Cycle Assess.</i> 18, 1325–1343 (2013). 18 98. Tobarra, M. A., López, L. A., Cadarso, M. A., Gómez, N. & Cazcarro, Its Seasonal Households' Consumption Good for the Nexus Carbon Water Footprint? the Spanish Fruits and Vegetables Case. <i>Environ. Sci. Technol.</i> 52, 12066–12077 (2018). 21 99. Theurl, M. C., Haberl, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long- range tomato production. <i>Agron. Sustain. Dev.</i> 34, 593–602 (2014). 23 100. Salendeeb, R., Font Vivanco, D., Al-Tabbaa, A. & zu Erngasseh, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Waste Manke</i>, 59, 442–450 (2017). 24 101. Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3– 18 (2013). 27 102. Hallström, E., Carlsson-Kanyama, A. & Rofresson, P. Environmental impact of dietary change: A systematic review. J. Clean, <i>Prod.</i> 91, 1–11 (2015). 29 103. Chane, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. J. Cleam, <i>Prod.</i> 91, 1–11 (2015). 21 104. Song, G., Li, M., Fullama-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate elimate change and benefit public health in China. <i>Sci. Tonal Environ</i>, <i>51</i>, 828–298 (2017). 31 104. Song, G., Li, M., Fullama-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate elimate change and benefit public health in China. <i>Sci. Tonal Environ</i>, <i>51</i>, 828–208 (2017). 31 10	17		95.	
 Weidor, J., Winning, K. (1994). In <i>Links</i>, Disk thoomsel, L. D. Your, and and the disk of general environmenal impact that the same foods produced within the UR? Int. J. Life Oxele ASKess. 18, 1325 (2013). Robert, M. A., López, L. A., Cadarso, M. A., Gómez, N. & Cazearro, I. Ts Seasonal Households' Consumption Good for the Nexus Carbon Water Footprint? the Spanish Fruits and Vegetables Case. <i>Environ. Sci. Technol.</i> 52, 12066–12077 (2018). P. Theurl, M. C., Haberl, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long-range tomato production. <i>Agron. Sustain. Dev.</i> 34, 593–602 (2014). Salendeeb, R., Font Vivanco, D., Al-Tabbaa, A. & <i>eu</i> Emgassch, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Waste Manke</i>, 59, 442–450 (2017). Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3–18 (2013). Hallström, E., Carlsson-Kanyama, A. & Börjesson, P. Environmental impact of dietary change: A systematic review. <i>J. Clean. Prod.</i> 91, 1–11 (2015). Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 91, 1–11 (2015). Song, G., Li, M., Fullana-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. <i>Sci. Total Environ.</i> 877, 289–298 (2017). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of randividual self-selected US dist. <i>Environ. Rev. Lett.</i> 13, (2018). Liu, G. & Mülller, D. B., Addreksing submainability in the aluminum industry: A critical review of life cycle assessments. <i>J. Clean. Prod.</i> 34, 008–117 (2012). Monsah, N. Y., TroBiborg, M., Kington, B., Aladers, I., & Hough, R. L. Greenhouse gas emi	19		96.	
 98. Tobarra, M. A., López, L. A., Cadarso, M. A., Gómez, N. & Cazcarro, I: Seasonal Households' Consumption Good for the Nexus Carbon/Water Footprint? the Spanish Fruits and Vegetables Case. Environ. Sci. Technol. 52, 12066–12077 (2018). 99. Theurl, M. C., Haberl, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long- range tomato production. Agron. Sustain. Dev. 34, 593-602 (2014). 91. Salemdeeb, R., Font Vivanco, D., Al-Tabbaa, A. & zu Ermipassen, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. Waste Mande, 59, 442-450 (2017). 101. Bellarby, J. et al. Livestock greenhouse gas emissions and mitigation potential in Europe. Glob. Chang. Biol. 19, 3– 18 (2013). 102. Hallström, E., Carlsson-Kanyama, A. & Börjesson, P. Environmental impact of dietary change: A systematic review. J. Clean. Prod. 91, 1–11 (2015). 103. Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. J. Clean. Prod. 91, 1–11 (2015). 104. Song, G., Li, M., Fullana-I-Palmer, P., Willamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. Sci. Total Environ. 577, 289–298 (2017). 105. Heller, M. C., Willins-Smith, A., Meyer, E., Keolcian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. Environ. Res. Lett. 13, (2018). 106. Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. J. Clean. Prod. 35, 109-117 (2012). 107. Dovjak, M., Markelj, J. & Kunte, R. Embodied global warming potential of different thermal insulation materials for industrial product: ARP J. Ling. Appl. Sci. 13, 2242–2249 (2018). 108. Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas e	22		97.	
 Theurl, M. C., Haberl, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long-range tomato production. <i>Agron. Sustain. Dev.</i> 34, 593–602 (2014). Salemdeeb, R., Font Vivanco, D., Al-Tabbaa, A. & <i>zu</i> Erngassen, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Waste Mandy.</i> 59, 442–450 (2017). Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3–18 (2013). Hallström, E., Carlsson-Kanyama, A. & Börgesson, P. Environmental impact of dietary change: A systematic review. <i>J. Clean. Prod.</i> 91, 1–11 (2015). Chune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 140, 766–783 (2017). Chue, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 140, 766–783 (2017). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). Liu, G. & Müller, D. B. Addressing sustainiability in the aluminum industry: A critical review of life cycle assessments. <i>J. Clean. Prod.</i> 35, 108–117 (2012). Jooyak, M., Markelj, J. & Komic, R. Embodied global warming potential of different thermal insulation materials for industrial products. <i>ARPN J. Eng. Appl. Sci.</i> 13, 2242–2249 (2018). Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy source: A review of lifecycle considerations. <i>Renev. Sustain. Energy Rev.</i> 29, 461–475 (2014). Rourke, J. W.O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission industrial products. <i>ARPN J. Eng. Appl. Sci.</i> 13, 2242–2249 (2018). Rourke, J. W.	24 25	19	98.	Good for the Nexus Carbon/Water Footprint? the Spanish Fruits and Vegetables Case. Environ. Sci. Technol. 52,
 Salemdeeb, R., Font Vivanco, D., Al-Tabbaa, A. & zu Ermgassen, E. K. H. J. A holistic approach to the environmental evaluation of food waste prevention. <i>Waste Manag.</i> 59, 442–450 (2017). Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3– 18 (2013). Hallström, E., Carlsson-Kanyama, A. & Börjesson, P. Environmental impact of dietary change: A systematic review. <i>J. Clean. Prod.</i> 91, 1–11 (2015). Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 140, 766-783 (2017). Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions and energy use associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). Heller, M., Markelj, J. & Kunie, R. Embodied global warming potential of different thermal insulation materials for industrial products. <i>ARPN J. Eng. Appl. Sci.</i> 13, 2242–2249 (2018). Amponsah, N. A., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. <i>Renew. Sustain. Energy Rev.</i> 39, 461–475 (2014). Rourke, J. M. O. & Seepersaf, C. C. The industrial protoxoltaic systems. <i>Proc. ASME</i> 2015 Int. <i>Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf 11</i> (2015). Ho Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation -	27	21	99.	Theurl, M. C., Haberl, H., Erb, K. H. & Lindenthal, T. Contrasted greenhouse gas emissions from local versus long-
 Bellarby, J. <i>et al.</i> Livestock greenhouse gas emissions and mitigation potential in Europe. <i>Glob. Chang. Biol.</i> 19, 3– 18 (2013). Hallström, E., Carlsson-Kanyama, A. & Börjesson, P. Environmental impact of dietary change: A systematic review. <i>J. Clean. Prod.</i> 91, 1–11 (2015). Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. <i>J. Clean. Prod.</i> 140, 766–783 (2017). Song, G., Li, M., Fullana-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. <i>Sci. Total Environ.</i> 577, 289–298 (2017). Song, G., Li, M., Fullana-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. <i>Sci. Total Environ.</i> 577, 289–298 (2017). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). Io6. Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. <i>J. Clean. Prod.</i> 35, 108–117 (2012). Io7. Dovjak, M., Markelj, J. & Kunič, R. Embodied global warming potential of different thermal insulation materials for industrial products. <i>ARPN J. Eng. Appl. Sci.</i> 13, 2242–2249 (2018). Io8. Amponsah, N. Y., Troldbore, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources. A review of lifecycle considerations. <i>Renew. Sustain. Energy Rev.</i> 39, 461–475 (2014). Io9. Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. <i>Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf. 1–11 (2015).</i> Matmodin, J. & Coroama, V. Assessing ICT	29 30		100.	
 Province State St	32		101.	
 Clune, S., Crossin, E. & Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. J. Clean. Prod. 140, 766-783 (2017). Song, G., Li, M., Fullana-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. Sci. Total Environ. 577, 289–298 (2017). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. Environ. Res. Lett. 13, (2018). Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production of individual self-selected US diets. Environ. Res. Lett. 13, (2018). Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. J. Clean. Prod. 35, 108–117 (2012). Dovjak, M., Markelj, J. & Kunič, R. Embodied global warming potential of different thermal insulation materials for industrial products. ARPN J. Eng. Appl. Sci. 13, 2242–2249 (2018). Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. Renew. Sustain. Energy Rev. 39, 461–475 (2014). Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf. 1–11 (2015). Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering ecosystem. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. Ecol. Econ. 145, 137–147 (2018). Hue, D. & Büchs, M. Household sharing for	34 35		102.	
 31 104. Song, G., Li, M., Fullana-i-Palmer, P., Williamson, D. & Wang, Y. Dietary changes to mitigate climate change and benefit public health in China. Sci. Total Environ. 577, 289–298 (2017). 33 105. Heller, M. C., Willits-Smith, A., Meyer, R., Keoleian, G. A. & Rose, D. Greenhouse gas emissions and energy use associated with production or individual self-selected US diets. Environ. Res. Lett. 13, (2018). 35 106. Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. J. Clean. Prod. 35, 108–117 (2012). 37 107. Dovjak, M., Markelj, J. & Kunič, R. Embodied global warming potential of different thermal insulation materials for industrial products. ARPN J. Eng. Appl. Sci. 13, 2242–2249 (2018). 39 108. Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. Renew. Sustain. Energy Rev. 39, 461–475 (2014). 41 109. Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf. 1–11 (2015). 44 110. Matmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering cosystem. 46 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. Ecol. Econ. 145, 137–147 (2018). 47 112. Vianova, D. & Bichs, M. Household sharing for carbon and energy reductions: the case of EU countries. Rev. 	37		103.	
 10. Hend, M. C., White Shind, M. Neyer K. Moster, S. D. Ordenhous gas emissions and chergy ase associated with production of individual self-selected US diets. <i>Environ. Res. Lett.</i> 13, (2018). 106. Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. <i>J. Clean. Prod.</i> 35, 108–117 (2012). 107. Dovjak, M., Markelj, J. & Kunič, R. Embodied global warming potential of different thermal insulation materials for industrial products. <i>ARPN J. Eng. Appl. Sci.</i> 13, 2242–2249 (2018). 108. Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. <i>Renew. Sustain. Energy Rev.</i> 39, 461–475 (2014). 109. Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. <i>Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf.</i> 1–11 (2015). 110. Matmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of supart metering Smart metering ecosystem. 46 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. <i>Ecol. Econ.</i> 145, 137–147 (2018). 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i> 	39 40		104.	
 44 35 106. Liu, G. & Müller, D. B. Addressing sustainability in the aluminum industry: A critical review of life cycle assessments. J. Clean. Prod. 35, 108–117 (2012). 46 37 107. Dovjak, M., Markelj, J. & Kunič, R. Embodied global warming potential of different thermal insulation materials for industrial products. ARPN J. Eng. Appl. Sci. 13, 2242–2249 (2018). 48 39 108. Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. Renew. Sustain. Energy Rev. 39, 461–475 (2014). 50 41 109. Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf. 1–11 (2015). 54 44 110. Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering ecosystem. 56 46 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. Ecol. Econ. 145, 137–147 (2018). 59 48 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. Rev. 	42		105.	
 industrial products. ARPN J. Eng. Appl. Sci. 13, 2242–2249 (2018). industrial products. ARPN J. Eng. Appl. Sci. 13, 2242–2249 (2018). 39 40 40 40 40 40 41 41 41 41 42 43 44 44 44 44 44 45 45 46 46 47 48 48 48 49 49 40 40 41 41 42 44 44 45 46 46 47 48 48 48 48 49 40 40 41 41 42 43 44 44 44 45 45 46 46 47 48 48 48 48 44 40 40 41 41 41 42 43 44 44 44 45 46 46 47 47 48 48 48 49 49 40 41 41 42 43 44 44 44 45 46 46 47 47 48 48 48 49 49 49 40 41 41 42 43 44 44 44 45 46 46 47 47 48 48 48 49 49 49 40 41 41 42 43 44 44 44 45 46 47 47 48 48 48 49 49 40 41 41 41 42 43 44 44 44 45 46 47 47 48 48 48 49 49 40 412 	44		106.	
 49 40 40 Amponsah, N. Y., Troldborg, M., Kington, B., Aalders, I. & Hough, R. L. Greenhouse gas emissions from renewable energy sources: A review of lifecycle considerations. <i>Renew. Sustain. Energy Rev.</i> 39, 461–475 (2014). 51 41 109. Rourke, J. M. O. & Seepersad, C. C. The importance of contextual factors in determining the greenhouse gas emission impacts of solar photovoltaic systems. <i>Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng. Conf.</i> 1–11 (2015). 54 44 110. Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering ecosystem. 56 46 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. <i>Ecol. Econ.</i> 145, 137–147 (2018). 59 48 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i> 	47		107.	
 42 emission impacts of solar photovoltaic systems. <i>Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng.</i> 43 <i>Conf.</i> 1–11 (2015). 44 110. Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering ecosystem. 46 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. <i>Ecol. Econ.</i> 145, 137–147 (2018). 48 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i> 	49		108.	
 44 110. Malmodin, J. & Coroama, V. Assessing ICT 's enabling effect through case study extrapolation – the example of smart metering Smart metering ecosystem. 46 46 47 111. Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. <i>Ecol. Econ.</i> 145, 137–147 (2018). 48 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i> 	51 52	42	109.	emission impacts of solar photovoltaic systems. Proc. ASME 2015 Int. Des. Eng. Tech. Conf. Comput. Inf. Eng.
 Fremstad, A., Underwood, A. & Zahran, S. The Environmental Impact of Sharing: Household and Urban Economies in CO2 Emissions. <i>Ecol. Econ.</i> 145, 137–147 (2018). 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i> 	54 55		110.	
48 112. Ivanova, D. & Büchs, M. Household sharing for carbon and energy reductions: the case of EU countries. <i>Rev.</i>	57		111.	
	59		112.	

Huebner, G. M. & Shipworth, D. All about size? - The potential of downsizing in reducing energy demand. Appl.

113.

2 Energy 186, 226-233 (2017). Ellsworth-Krebs, K., Reid, L. & Hunter, C. J. Home Comfort and "Peak Household": Implications for Energy 114. Demand. Housing, Theory Soc. 00, 1-20 (2019). Cheng, M. et al. The sharing economy and sustainability – assessing Airbnb's direct, indirect and induced carbon 115. footprint in Sydney. J. Sustain. Tour. 0, 1-17 (2020). Vita, G. et al. The Environmental Impact of Green Consumption and Sufficiency Lifestyles Scenarios in Europe: 116. Connecting Local Sustainability Visions to Global Consequences. Ecol. Econ. 164, (2019). 117. Evidence, C. for E. Collaboration for Environmental Evidence Synthesis Assessment Tool (CEESAT) criteria and scoring guidelines for reliability of evidence For details on appropriate use and limitations of this tool please refer to Woodcock et al (2013) Biological Conser. (2014). Guan, D. et al. Determinants of stagnating carbon intensity in China. Nat. Clim. Chang. 4, 1017-1023 (2014). 118. Chen, Z. M. et al. Consumption-based greenhouse gas emissions accounting with capital stock change highlights 119. dynamics of fast-developing countries. Nat. Commun. 9, (2018). 120. Tong, D. et al. Committed emissions from existing energy infrastructure jeopardize 1.5 °C climate target. Nature (2019). doi:10.1038/s41586-019-1364-3 González-García, S., Esteve-Llorens, X., Moreira, M. T. & Feijoo, G. Carbon footprint and nutritional quality of 121. different human dietary choices. Sci. Total Environ. 644, 77-94 (2018). Heller, M. C. & Keoleian, G. A. Greenhouse Gas Emission Estimates of U.S. Dietary Choices and Food Loss. J. Ind. 122. Ecol. 19, 391-401 (2015). Werner, L. B., Flysjö, A. & Tholstrup, T. Greenhouse gas emissions of realistic dietary choices in Denmark: The 123. carbon footprint and nutritional value of dairy products. Food Nutr. Res. 58, 1-16 (2014). Wolfram, P. & Hertwich, E. Representing vehicle-technological opportunities in integrated energy modeling. 124. Transp. Res. Part D Transp. Environ. 73, 76–86 (2019). Seto, K. C. et al. Carbon lock-in: Types, causes, and policy implications. Annu. Rev. Environ. Resour. 41, 425-452 125. (2016). Nyborg, K. et al. Social norms as solutions: Policies may influence large-scale behavioral tipping. Science (80-.). 126. 354, 42-43 (2016). 127. Di Giulio, A. & Fuchs, D. Sustainable consumption corridors: Concept, objections, and responses. Gaia 23, 184-192 (2014). 128. Haddaway, N. R., Macura, B., Whaley, P. & Pullin, A. S. ROSES flow diagram for systematic reviews. Version 1.0. (2017). doi:10.6084/m9.figshare.5897389